

Thermal and chemical enhanced recovery of heavy chlorinated organic compounds in saturated porous media: 1D cell drainage-imbibition experiments

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1	Thermal and chemical enhanced recovery of heavy chlorinated organic
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4	S. Colombano ^{1*} , H. Davarzani ¹ , E.D van Hullebusch ² , D. Huguenot ³ , D. Guyonnet ¹ , J. Deparis ¹ , I.
5	Ignatiadis ¹
6	
7	¹ Bureau de Recherches Géologiques et Minières (BRGM), France
8	² Université de Paris, Institut de Physique du Globe de Paris, CNRS, F-75005 Paris, France
9	³ Laboratoire Géomatériaux et Environnement, Université Paris-Est Marne-la-Vallée, France
10	
11	*corresponding author: Stéfan Colombano: s.colombano@brgm.fr

2 Abstract

3

4 Chemical and thermal enhanced recovery of pure heavy chlorinated organic compounds (DNAPL; dense non-aqueous phase liquids) were investigated by using lab-scale 1D cells. 5 Temperature was increased to reduce DNAPL viscosity (and hence increase its mobility), 6 7 while surfactant was added to decrease capillary forces involved in the entrapment of DNAPL in porous media. Laboratory scale experiments, based on mass balance and indirect 8 monitoring methods (*i.e.*, permittivity, electrical resistivity and optical density), were 9 conducted to quantify the effects of these enhancements. Heating the DNAPL up to 50 °C 10 decreased its viscosity by a factor of two. The addition of a surfactant; *i.e.*, Sodium Dodecyl 11 Benzene Sulfonate (SDBS), at its Critical Micelle Concentration (to prevent DNAPL 12 solubilization), decreased interfacial tensions by a factor of 12. Drainage-imbibition 13 experiments performed in 1D cells provided retention curves (capillary pressure as a function 14 15 of water saturation) of a two-phase (DNAPL-water) system in experimental glass bead porous media. The observed reduction of residual saturation (Sm) obtained with SDBS was 28% for 16 0.5 mm-diameter glass beads (GB) and 46% for 0.1 mm GB. No significant decrease in S_m 17 18 was observed with thermal enhancement. The van Genuchten - Mualem model was found to satisfactorily reproduce the measured retention curves. Indirect measurements of water 19 20 saturations (S_w) showed that: i. measured permittivities were very close to values modeled 21 with the Complex Refractive Index Model (CRIM); ii. Archie's Law was less successful in reproducing measured electrical resistivities; iii. optical densities provide accurate estimations 22 of S_w. At field scale, the combined monitoring of electrical resistivity (which provides a 23 24 global picture) and permittivity (which yields locally precise but spatially limited information) is expected to significantly improve the collection of information on residual 25 26 saturations S_{rn}.

28 Keywords

29

30 Dense Non-Aqueous Phase Liquids, chemical enhancement, thermal enhancement,

31 permittivity, electrical resistivity, optical density

32

34 1. Introduction

35 Chlorinated Organic Compounds (COCs) have been produced in large quantities since the middle of the 20th century, mostly for chemical, pharmaceutical and agricultural applications [Cohen and 36 Mercer (1993); Kueper et al. (2003)]. Due to their high hydrophobicity and their density higher than 37 that of water, COCs may infiltrate aquifers vertically and form DNAPL (Dense Non-Aqueous Phase 38 39 Liquid) pools [Schwille (1988); Cohen and Mercer (1993)]. Their dispersion (by solubilization and volatilization) from the pollution source zone may generate large contaminant plumes leading to 40 severe groundwater contamination [Stupp and Paus (1999)]. These particularly toxic pollutants, may 41 42 permanently affect soil, indoor air and groundwater quality [ADEME and Ernst & Young (2014); NIEHS (2015); IARC (2018)]. 43

44 Treating such contaminant sources is therefore a priority in order to avoid the migration of DNAPL
45 to groundwater over very long periods of time, leading to a buildup of contaminant plume
46 concentrations and plume spreading [Sale (2001); Stroo et al. (2003); Falta et al. (2005a); Falta et al.
47 (2005b); McDade et al. (2005); Newell and Adamson (2005); McGuire et al. (2006); Huang et al.
48 (2015)].

49 The recovery of COC free-phase (mobile) product is typically performed using pump and treat 50 methods. Due to COC high density, low solubility and strong interfacial tension with water, pump and treat operations are time-consuming (e.g., commonly in excess of 30 years). They are also not very 51 52 effective in the long term due to the slow release from the residual saturation and the low remediation yield [Mackay and Cherry (1989); Travis and Doty (1990); Berglund and Cvetkovic (1995); Pankow 53 54 and Cherry (1996); Falta et al. (2005b); McDade et al. (2005); McGuire et al. (2006); Harkness and 55 Konzuk (2014)]. Typical recovery yields following pumping do not exceed 60% [ITRC (2002); Stroo 56 et al. (2012)]. DNAPL mobilization only starts when the sum of viscous and buoyancy forces exceed 57 capillary forces within the contaminated medium [Pennell et al. (1996); Duffield and Ramamurthy (2003)]. 58

59 The DNAPL studied herein was sampled from the subsurface of the Tavaux site; a large chloralkali60 chemical plant located in the center-east of France. It is a mixture composed mainly of heavy

chlorinated organic compounds (weight percent; wt%): hexachlorobutadiene-HCBD (58%), hexachloroethane-HCA (14%), perchloroethylene-PCE (8%), and pentachlorobenzene (3.5%), carbon tetrachloride (4%), trichloroethylene-TCE (2%), hexachlorobenzene (1%) [Cazaux et al. (2014)]. Due to the accidental release of DNAPL, the subsurface of the Tavaux site is contaminated, with locally pure DNAPL present below the water table. Investigations aim therefore at optimizing the recovery of this free phase, either through thermal (heating) or chemical (surfactants) enhancements.

67 Surfactants addition aims to: a) decrease the interfacial tension (IFT) between the organic phase and water and b) increase NAPL solubility in water through micelle generation. Dissolution of 68 contaminants into micelles arises when surfactant concentrations exceed the Critical Micelle 69 70 Concentration (CMC) value in the pores [Rosen (1989)]. While the chlorinated compounds that are the 71 most documented in the literature are PCE and TCE, the DNAPL investigated in the present study does not contain more than 15% (wt%) of these compounds. Also, there is a paucity of literature 72 73 regarding the main compounds encountered on the Tavaux DNAPL (*i.e.*, HCBD and HCA). Nonionic surfactants are effective for remediating chlorinated solvents such as PCE and TCE because of their 74 75 ability to reduce the IFT and increase pollutant solubility [Taylor et al. (2001); Zhong et al. (2003); Zhao et al. (2006); Suchomel et al. (2007); Harendra and Vipulanandan (2011); Atteia et al. (2013); 76 77 Pennell et al. (2014)]. Several surfactants (i.e., Triton X-100, Tween 80, Aerosol MA-80 and SDBS) 78 are able to solubilize and reduce the IFT for the TCE-water and PCE-water systems. For TCE, IFT has 79 been reduced from 35.2 to 0.2 mN.m⁻¹ with Aerosol-MA-80 [Dwarakanath et al. (1999)] and from 35.2 to 10.4 mN.m⁻¹ with Tween 80 [Suchomel et al. (2007)]. For PCE, the system's IFT can be reduced by 80 the Aerosol family of surfactants: from 47.8 to less than 0.01 mN.m⁻¹ [Dwarakanath et al. (1999); 81 Sabatini et al. (2000); Childs et al. (2004)], or by Triton X-100 and Tween 80 [Taylor et al. (2001); 82 83 Harendra and Vipulanandan (2011)]. In this work, the surfactants were introduced at concentrations 84 well below their respective CMC concentrations, in order to avoid DNAPL dissolution and hence the 85 need for costly water treatment technologies [Sabatini et al. (1998); Ahn et al. (2008); Atteia et al. (2013); Maire et al. (2018)]. 86

Regarding temperature enhancement, it has been shown that the viscosity of a chlorinated solvent is
generally reduced by 1% per degree Celsius incremental increase [Davis (1997)]. [Sleep and Ma

89 (1997)] observed a significant viscosity reduction with increasing temperature in the case of PCE. 90 Such viscosity reduction increases DNAPL mobility and therefore decreases residual saturations. She 91 and Sleep (1998) reported that increasing the temperature decreased residual PCE [She and Sleep 92 (1998)]. Kong (2004) also demonstrated during imbibition-drainage experiments that residual coal tar 93 saturations reduced when temperatures were increased from 22 to 50 °C (respectively $S_{rn} = 31.4 \%$ to 94 27.6 % in F-70 sand and $S_{rn} = 32.8 \%$ to 21.8 % in 20-30 mesh Ottawa sand) [Kong (2004)]. Indirect 95 monitoring of these residual saturations can be performed using several methods.

Permittivity measurements using Time-Domain Reflectometry (TDR) probes have served as 96 97 indicators of Dense Non-Aqueous Phase Liquid (NAPL) saturation (S_n) [Redman et al. (1991); Redman and DeRyck (1994); Kueper et al. (1993); Brewster et al. (1995)]. PCE samples spiked at 98 controlled values of saturation were monitored by TDR, with results showing a good correlation 99 100 between PCE saturation and permittivity interpreted using a mixing model [Redman and DeRyck (1994)]. The most commonly used model for correlating dielectric permittivity to saturation levels in a 101 soil made of a mixture of particles, water and air, is the Complex Refractive Index Model (CRIM) 102 103 [Birchak et al. (1974); Roth et al. (1990); Endres and Knight (1992); Ajo-Franklin et al. (2004)]. 104 Electrical resistivity is another parameter used as a means to estimate DNAPL saturation levels. Interpretation is performed using Archie's Law, which describes how resistivity depends on porosity if 105 106 ionic conduction in the pore fluid dominates other conduction mechanisms in the rocks [Archie 107 (1942)]. Resistivity is described as the sum of a bulk conductivity term and a surface conductivity 108 term, both of which are saturation-dependent [Revil (2012)].

109 Finally, the Light Reflection Method (LRM) was used to calibrate NAPL saturation with optical 110 density [Alazaiza et al. (2016)]. Schincariol and Schwartz (1990) and Schincariol et al. (1993) published the first studies showing a correlation between a reflected optical image and dyed NaCl 111 112 concentrations during miscible experiments [Schincariol and Schwartz (1990); Schincariol et al. (1993)]. Flores et al. (2011) quantified the residual saturation of LNAPL using Sudan III dyed 113 (synthetic organic compounds) with brilliant blue FCF and Simplified Image Analysis Method 114 (SIAM). The difference between mass balance and image interpretation was only 4.7% [Flores et al. 115 (2011)]. O'Carroll et al. (2004) used LRM to estimate the saturation of tetrachloroethylene (PCE) in a 116

117 flow chamber and observed a linear relationship between the hue and PCE saturation (with a 118 correlation R^2 value of 0.91) [O'Carroll et al. (2004)]. Luciano *et al.* (2010) used mass balance 119 calculations to validate the suitability of LRM, with a difference between mass balance and image 120 interpretation of 6% [Luciano et al. (2010)]. Unlike previous authors, the approach adopted herein 121 relied on a calibration based on a superior number (4) of water saturation calibration points ($S_w = 0\%$, 122 100%, 1-S_{rn}, and S_{rw}).

The purpose of this study was to assess the potential of chemical and thermal enhancements for improving DNAPL recovery during pumping. Investigations were performed at lab scale with controlled 1D cells, a DNAPL mixture collected in the field and porous media made of glass beads. This experimental setup helped avoid potential artefact effects associated with porous medium heterogeneity, while focusing on primary mechanisms and parameters (*e.g.*, interfacial tension, micellar concentrations, ...) and on the indirect measurement of DNAPL saturation levels using three different methods (permittivity, electrical resistivity and optical density).

130

131 2. Materials and methods

The DNAPL used for the experiments was sampled at the Tavaux site (June 25, 2014) and stored in a cold room (at 4 °C). DNAPL was filtered just before its use (EMD Millipore, 0.45 μ m). The water used for all experiments was BRGM tap water degassed using an ultrasound tank (VWR Ultrasonic Cleaner - USC500D: 60 °C, 45 Hz, 60 min). This tap water is extracted from a groundwater well that shows very stable chemical characteristics.

The experiments with thermal enhancement were carried out in the 20-50 °C temperature range (in order to avoid DNAPL volatilization). Four surfactants were tested at their CMC: 2 non-ionic (Triton X-100: 4-(1,1,3,3-Tetramethylbutyl)phenyl-polyethylene glycol and Tween 80: sorbitan mono-9octadecenoate poly(oxy-1,2-ethanediyl)) and 2 anionic (Aerosol MA-80: sodium 1,4-dicyclohexyl sulphonatosuccinate and SDBS: sodium dodecyl benzene sulfonate). All surfactants were supplied by Sigma Aldrich (laboratory grade), except for the Aerosol MA-80, which was supplied by Cytec.

143

144 2.1 Characterization of fluid physical properties

All measurements presented herein were carried out in triplicate. Viscosity was first measured by Rheonova Laboratory (Grenoble, France) using a Discovery Hybrid Rheometer - DHR 3 (TA Instruments) and then at BRGM with a Haake Mars III rheometer from Thermo Fischer Scientific. Measurements were performed at atmospheric pressure and at 10, 15, 20, 30, 45 and 60 °C. Temperature was controlled through the Peltier effect. The temperature precision was 0.5 °C.

150 The IFT and contact angle between the DNAPL and water were measured at the Navier laboratory (Champs-sur-Marne, France) with a Tracker-S tensiometer (Teclis Scientific) and also at BRGM with 151 a Drop Shape Analyser tensiometer DSA-100 (Krüss). The Tracker-S and the DSA-100 are automated 152 153 drop tensiometers that can measure variations in IFT over time. The instrument can also be used to measure the contact angle between a liquid and a solid. Using optional lenses, the accuracy of IFT 154 measurement can be increased to 0.1 mN.m⁻¹. The IFT and contact angle for DNAPL-water were 155 measured at different temperatures (10 to 60 °C) and with surfactants at different concentrations. The 156 temperature was stabilized with a thermostatically controlled water bath. Product densities were 157 158 measured by first weighing three different 100 mL flasks and then weighing, for each temperature, the product was weighed in the closed vial while adjusting the volume required to fill the flasks. 159

160

161 2.2 Solubilization

DNAPL solubility measurements were performed using the experimental protocol of Rodrigues *et al.* (2017). The same protocol was applied for solubilization experiments in the presence of surfactants. Four surfactant concentrations were studied, according to their respective CMC value, from 0.25×CMC to 8×CMC. All measurements were performed in triplicate at 20 °C [Rodrigues et al. (2017)].

167 COC concentrations in water were analyzed by gas chromatography with flame ionization detector
168 (Varian 3800) and head space injection (Agilent 7697A). The column used was an Agilent CP-SIL 5
169 CB (semi-capillary column, fused silica tube, length: 50 m, internal diameter: 0.53 mm, active phase:
170 polydimethylsiloxane, carrier gas: helium).

171

172 2.3 Drainage-imbibition experiments in 1D cells

The main objective of the experimental study was to characterize the soil parameters in two-phase 173 174 flow conditions (DNAPL and water) in a porous medium and under different conditions (thermal or chemical enhancement). To achieve this objective, the laboratory work was dedicated to performing 175 drainage-imbibition experiments in 1D cells (filled with glass beads; GB) that were used to generate 176 the capillary pressure-saturation curves in static flow conditions. The 1D cells had an internal diameter 177 178 of 5.8 cm and a height of 5.56 cm (Figure 1). Two sizes of GB were used to fill the cells in order to 179 reproduce the hydraulic conductivities measured at the site: 0.10 ± 0.02 mm and 0.5 ± 0.1 mm GB. The cells were carefully filled with GB and vibrated. The mass of GB porous medium inside the cells 180 was kept constant in order to ensure similar porous medium compaction and porosity charateristics. 181 The GB measured permeabilities were 1.30×10^{-10} m² and 6.73×10^{-12} m² for 0.5 mm GB and 0.1 mm 182 GB, respectively. At both extremities, the main cells were connected via two tubes to two graduated 183 "reservoir" columns (3.5 cm internal diameter and 41.8 cm high). These two reservoirs (containing 184 respectively water and DNAPL) stored the fluids and allowed to control the pressure heads at the 185 186 lower and upper parts of the cell. The cells and the columns were made of PolyVinyliDene Fluoride (PVDF) and were connected by thermo-scientific Nalgene 8001-1014 | 180 metric clear PolyVinyl 187 Chloride (PVC) tubes with 1 cm internal diameter. The volumes of DNAPL and water in the porous 188 189 media were calculated by volume balance. The addition of surfactants and the heating up to 50 °C 190 were carried out at the end of the drainage periods in order to start with the same values of S_{rw} as for the experiments without enhancement. The experimental set-up was also constantly monitored by 191 192 geophysics and TDR probes. The number of drainage-imbibition experiments performed with 0.5 and 193 0.1 mm GB were respectively: 5 and 6 without enhancement and with membrane, 3 and 4 without 194 enhancement and without membrane, 8 and 6 with surfactants, 2 and 2 with thermal enhancement.

The curves, which were afterwards calibrated using the van Genuchten-Mualem model (VGM), provide key parameters such as the irreducible water saturation (S_{rw}), the residual pollutant saturation (S_m), the entry pressure, and the calibration parameters α and n which characterize the capillary forces and the heterogeneity of the porous medium, respectively (Eq. 1 and Eq. 2) [Mualem (1976); van Genuchten (1980)].

$$S_{ew} = \frac{1}{[1 + (\alpha h_c)^n]^m}$$
 Eq. 1

$$S_{ew} = \frac{S_w - S_{rw}}{1 - S_{rw} - S_{rn}}$$
 Eq. 2

where m = 1-1/n, $S_{ew}(-)$ is the effective water saturation, h_c (m) is the capillary pressure head, α (m⁻¹) is a fitting parameter that is inversely proportional to the non-wetting fluid entry pressure value, and n (-) is the width of pore-size distribution.

203 The experimental data were fitted to the VGM capillary pressure-saturation function using the204 solver provided in Excel by minimizing the Sum of Squared Errors (SSE).

205

206 2.4 Monitoring of the drainage-imbibition experiments with dielectric permittivity

Dielectric permittivities were monitored using TDR probes (Decagon Devices 5TE 40567) with a 70 MHz frequency. Data acquisition frequency was 2 signals per minute. In addition to permittivity (-), these probes monitor volumetric water content-VWC (m³.m⁻³), temperature (°C), and bulk electrical conductivity-EC (dS.m⁻¹). They were connected to a Campbell Scientific CR1000 (4M) data logger to acquire temperature and permittivity data. The permittivity values were corrected relative to the reference values measured in air and water using the following equation (Eq. 3) [Kargas and Soulis (2012)]:

$$\varepsilon_{corrected} = \frac{\varepsilon_{w-\text{theoretical}}(\varepsilon_{measured} - \varepsilon_{air})}{(\varepsilon_{w} - \varepsilon_{air})} + \varepsilon_{air-\text{theoretical}}$$
Eq. 3

where $\varepsilon_{w-\text{theoretical}}$ (-) is the relative effective permittivity of pure water ($\varepsilon_{w-\text{theoretical}}$ = 80 at 20 °C), $\varepsilon_{measured}$ (-) is the relative effective permittivity of the medium measured with the TDR probe during the experiments, ε_{air} (-) is the relative effective permittivity of air measured at the beginning of the experiment, ε_w (-) is the relative effective permittivity of tap water measured at the beginning of the experiment, ε_w (-) is the relative effective permittivity of tap water measured at the beginning of the experiment, and $\varepsilon_{air-theoretical}$ (-) is the relative effective permittivity of pure air ($\varepsilon_{air-theoretical}$ = 1).

219

220 2.5 Monitoring of the drainage-imbibition experiments with electrical resistivity

Electrical resistivity was monitored using unpolarizable potential electrodes for voltage
 measurement, metallic current electrodes for current injection, a resistivity meter and data acquisition

software. To limit electrode polarization phenomena, which can add noise to the measurements, potential electrodes must be unpolarizable [Dahlin (2000)]. Cu/CuSO₄ electrodes were used and the method developed by Noel [Noel (2014)]. These electrodes were made by mixing milli-Q water 72.2%, CuSO₄ 26% and Gelatin 1.7% and heated the mixture to \approx 80 °C for 45 minutes using a shaking heating plate. The resulting electrodes are unpolarizable, *i.e.*, they do not polarize the ground and therefore the electrode action on the potential measured can be considered negligible.

The metallic current electrodes were made of nickel-cobalt alloy (MP35N). The resistivity meter was SIP LAB IV and data acquisition software was SIP LAB IV. Resistivity was measured between 0.1 to 20 000 Hz but analysed at 1.4 Hz. This frequency was chosen because it is close to that used in the field [Chambers et al. (2004); Constable and Srnka (2007); Han et al. (2015); Deparis et al. (2019)]. The Wenner electrodes configuration was applied, as it well suited for cell experiments [Noel (2014)].

The values measured were resistance values, R_e , which were transformed into apparent electrical resistivity values, ρ_c , using the geometric coefficient, K_g (Eq. 4 and Eq. 5) [Reynolds (2011)] :

$$K_g = \frac{R_0}{R_e}$$
 Eq. 4

$$\rho_c = R_e K_g$$
 Eq. 5

237 where K_g (m) is the geometric coefficient, and R_0 (Ω .m) is the initial estimated resistance value.

The conductivity of the tap water, measured at the beginning of each experiment, made it possible to determine R_o and to calculate K_g (using Eq. 4). The value of K_g was then used throughout the experiment to transform R_e into ρ_c .

Experiments were carried out to quantify the impact of chemical and thermal enhancements on permittivities and resistivities: drainage experiments were first performed with only tap water and GB by (i) introducing tap water with surfactants (at their CMC values) and (ii) increasing temperature. Experiments were then performed in order to quantify the effects of temperature on the permittivity and resistivity of DNAPL, water, DNAPL+GB, and water+GB systems. In these experiments, temperature was increased stepwise from 10 to 50 °C. Permittivity and resistivity were noted once the temperature of each step had stabilized. The correlation curves helped to correct the permittivity andresistivity values with temperature variations.

249

250

2.6 Monitoring of the drainage-imbibition experiments with image interpretation

The optical imaging method used was the light reflection method (LRM), with a Nikon® D810 251 with NIKKOR LENS 105 (Nikon®) digital camera which has a high resolution of 34 Mega Pixels. The 252 253 image resolution calculated from Fiji (open source image processing package based on ImageJ) for these experiments was approximately 0.003 mm².pixel⁻¹, depending on the image size [pixels×pixels] and the 254 distance between the camera and the object of interest. The following camera set up was used: aperture 255 = 1/200 s, ISO = 100 and the shutter = f/16. This set up remained the same during all experiments. The 256 257 Capture One® software was used to take photographs without touching the camera. All pictures were 258 acquired on RAW format (.raw) to save the full data information.

259 For the calibration, the drainage-imbibition experiments were conducted in the same way as with 260 1D cells but this time with a Hele-Shaw cell (to prevent light reflection). The dimensions of this cell 261 were: height = 5.00 cm, length = 5.00 cm and width = 2.00 cm. The cell was made of PVDF (as with 262 the 1D cells), to ensure high chemical resistance to the pure pollutant. The cell was composed of two transparent glass faces to allow photographing. The experiments were performed in a dark room and 263 264 the light source was provided by two 2×300 W floodlights (Broncolor®). The camera was always placed in the same position for all experiments. A color scale was placed beside the cell to calibrate 265 the differences, albeit small, between the lighting used for the various experiments. 266

First, a global Area Of Interest (AOI) was defined to obtain the mean grey value necessary to associate with S_w. Then each picture was converted into 8-bit format. Finally, the mean grey values were calculated using Fiji software, for different AOI centered in the middle of the picture. Fiji software was chosen because it is easy to apply a threshold that depends on the pixel intensity value. The threshold depends on the calibration curve fluid saturation versus intensity. Using Fiji, it is possible to determine the area associated with the threshold. From this information (DNAPL saturation and its own area), one can compute the DNAPL volumes present using the following expression (Eq.

274 6).

$$V_{DNAPL} = S_{DNAPL} A_{DNAPL} l \emptyset$$
 Eq. 6

where V_{DNAPL} (m³) is the volume of DNAPL for a given saturation, S_{DNAPL} (-) is the DNAPL saturation for a given threshold on Fiji, A_{DNAPL} (m²) is the area associated of the given saturation, 1 (m) is the length between the front wall and the back wall of the cell, and \emptyset (-) is the porosity.

The image data analysis was performed according to the following main steps: a) convert the picture into 8 bit format to obtain 256 shades of gray; b) set up the scale; c) set up the I_0 (Initial luminous intensity), which was equal to white on the gray scale; d) define the AOI; e) set up the contrast to optimize the black pixels; f) compute the layers of DNAPL saturation present on the AOI, based on the linear relation derived from the calibration experiment.

283

284 **3.** Theory

The parameters presented in section 2 provide information regarding the potential for mobilizing DNAPL in a porous medium, using various relationships or models. For example, Pennell *et al.* (1996) used a method to estimate DNAPL mobilization in the porous medium with capillary number (N_{ca}), Bond number (N_B) and total trapping number (N_T) (Eq. 7 to Eq. 9).

$$N_{ca} = \frac{v_w \mu_w}{\sigma \cos \theta} \qquad \qquad \text{Eq. 7}$$

$$N_{\rm B} = \frac{g k k_{\rm rw} \Delta \rho}{\sigma cos \theta}$$
 Eq. 8

$$N_{T} = \sqrt{N_{ca}^{2} + 2N_{ca}N_{B}sin\alpha_{N_{T}} + N_{B}^{2}}$$
 Eq. 9

where v_w (m.s⁻¹) is the Darcy velocity of the wetting phase (upward direction is considered positive), μ_w (Pa.s) is the dynamic viscosity of the wetting phase, $\Delta \rho$ (kg.m⁻³) is the difference of densities between the wetting and non-wetting phase (= ρ_w - ρ_n), and α_{N_T} (°) is the angle between the direction of flow and the horizontal direction.

293 The mobility ratio (m_r) can be calculated as follows (Eq. 10) [Lenormand et al. (1988); Dullien
294 (1992)]:

$$m_r = \frac{k_{2r}\mu_1}{k_{1r}\mu_2}$$
 Eq. 10

where m_r (-) is the mobility ratio, k_{1r} (-) is the relative permeability for the displaced phase, k_{2r} (-) is the relative permeability for the displacing phase, μ_1 (Pa.s) is the fluid dynamic viscosity for the displaced phase, and μ_2 (Pa.s) is the fluid dynamic viscosity for the displacing phase.

Regarding the monitoring of permittivities to characterize saturations, the CRIM model has been
developed for multiphase systems (Eq. 11) [Birchak et al. (1974); Roth et al. (1990); Endres and
Knight (1992)]:

$$\varepsilon = \left[\sum_{i=1}^{N} v_i \varepsilon_i^{\alpha_{\varepsilon}}\right]^{1/\alpha_{\varepsilon}}$$
Eq. 11

301 where ε (-) is the relative effective permittivity of the mixture, ε_i (-) is the relative effective 302 permittivity of the i phase, v_i (-) is the volume of the i phase, and α_{ε} (-) is an empirical constant 303 related to the geometry of the grains and their spatial distribution.

Ajo-Franklin *et al.* (2004) report correlations higher than 97% using the CRIM model with TCE in the saturated zone [Ajo-Franklin et al. (2004)]. For a three-phase mineral/water/NAPL mixture and assuming that $\alpha_c = 0.5$, the CRIM equation becomes (Eq. 12):

$$\varepsilon = \left[\emptyset \left(S_w \sqrt{\varepsilon_w} + S_n \sqrt{\varepsilon_n} \right) + (1 - \emptyset) \sqrt{\varepsilon_s} \right]^2$$
 Eq. 12

307 where S_w (-) is the water (wetting fluid) saturation, ε_w (-) is the relative effective permittivity of water, 308 S_n (-) is the DNAPL (non-wetting fluid) saturation, ε_n (-) is the relative effective permittivity of 309 NAPL, and ε_s (-) is the relative effective permittivity of soil particles.

For electrical resistivity interpretation, Archie's Law is described as follows (Eq. 13) [Archie(1942)]:

$$\rho_c = \rho_{c,w} a_c \phi^{-m_c} \qquad \qquad \text{Eq. 13}$$

where ρ_c (Ω .m) is the real effective electrical resistivity of the bulk, $\rho_{c,w}$ (Ω .m) is the real effective electrical resistivity of the fluid at temperature T, a_c (-) is an empirical parameter (typically equal to 1 but that can vary from <1 for intergranular porosity to > 1 for joint porosity), and m_c (-) is the cementing factor (also an empirical parameter usually approximately equal to 2, but that car vary from 1.2 for unconsolidated sediments to 3.5 for crystalline rocks). Glover (2010) proposed, on the basis of Archie's law, the following equations to calculate S_w for a twophase system, where neglecting the conductivity of GB is neglected as it is very low ($\sigma_{c,GB} = 10^{-20} \text{ S.m}^{-1}$) (Eq. 14 to Eq. 17) [Glover (2010)]:

$$\sigma_{c,bulk} = \sigma_{c,DNAPL} [\emptyset(1 - S_w)]^{m_1} + \sigma_{c,water} [\emptyset(S_w)]^{m_2}$$
Eq. 14

$$\left(-\frac{\phi_1^2}{2}\right)m_2^2 + \left(\phi_1 + \frac{\phi_1^2}{2}\right)m_2 - \phi_1^{m_1} = 0$$
 Eq. 15

$$m_{2} = \frac{-\left(\emptyset_{1} + \frac{\emptyset_{1}^{2}}{2}\right) \pm \left[\left(\emptyset_{1} + \frac{\emptyset_{1}^{2}}{2}\right)^{2} - 4\left(-\frac{\emptyset_{1}^{2}}{2}\right)\left(-\emptyset_{1}^{m_{1}}\right)\right]^{\frac{1}{2}}}{-\emptyset_{1}^{2}}$$
Eq. 16

$$m_2 = \frac{-(4\phi_1 + 2\phi_1^2) \pm (4\phi_1^2 + 4\phi_1^3 + \phi_1^4 - 8\phi_1^2\phi_1^{m_1})^{\frac{1}{2}}}{-4\phi_1^2}$$
 Eq. 17

where $\sigma_{c,bulk} = \sigma_c (S.m^{-1})$ is the real effective electrical conductivity of the bulk, $\sigma_{c,DNAPL} (S.m^{-1})$ is the real effective electrical conductivity of DNAPL, m_1 (-) is the cementation exponent of DNAPL phase, $\sigma_{c,water} (S.m^{-1})$ is the real effective electrical conductivity of water, m_2 (-) is the cementation exponent of water phase, and $\phi_1 = \phi(1 - S_w); \phi_2 = \phi(S_w)$.

For image analysis, the optical density of reflected light, D_r, can be defined as (Eq. 18 and Eq. 19)
[Stimson (1974); Schincariol et al. (1993); Flores et al. (2011)]:

$$D_r = -\lg\left(\rho_t\right) \qquad \qquad \text{Eq. 18}$$

$$\rho_t = \frac{I_r}{I_0}$$
 Eq. 19

where: $D_r = OD$ (-) is the optical density of reflected light, ρ_t (-) is the ratio of reflected/initial luminous intensity, I_r (-) is the reflected luminous intensity, and I_0 (-) is the initial luminous intensity.

- 329 4. Results and discussion
- 330 4.1 Fluids characterization
- 331 4.1.1 Dynamic viscosity

332 While chlorinated solvents are typically more fluid than water (with viscosities less than 1 mPa.s)

[Sleep and Ma (1997)], the dynamic viscosity of the DNAPL studied herein was much higher (5 mPa.s

at groundwater temperature, *i.e.* 12 °C). The viscosity of a chlorinated solvent is generally reduced by 334 1% per degree Celsius [Davis (1997)]. The dynamic viscosity of the DNAPL studied here reduced by 335 336 57% when the temperature rose from 10 to 60 °C, which is in agreement with the literature (Supplementary materials, Figure A-1). As shown in this figure, the change in dynamic viscosity as a 337 function of temperature fitted a second order polynomial curve. For water, the reduction in dynamic 338 viscosity was 54% for the same temperatures. The μ_w/μ_{nw} ratio ranged from 0.23 to 0.19 for 339 340 temperatures of 10 °C and 60 °C, respectively. This 15% reduction in viscosity ratios may contribute to improve DNAPL mobility. 341

342

343 4.1.2 Interfacial tension and contact angle between DNAPL and water

The IFT, σ , was measured as 11.15± 0.05 mN.m⁻¹ at 20°C (Supplementary materials, Figure A-2). 344 This value was lower than that of pure TCE or PCE in distilled water (respectively 36.9 and 45.9 345 mN.m⁻¹) [Andersson et al. (2014)]. The results for reduced interfacial tensions with added surfactants 346 347 were very contrasted. Three trends were observed: a moderate effect for Tween 80, a substantial effect 348 for relatively high CMC of Triton X-100 and Aerosol MA-80, and a very substantial beneficial effect for low surfactant concentrations of SDBS. More precisely, we found that the reduction in IFT for 349 Tween 80 was 52% with the concentration of 64×CMC ($\sigma_{\text{final}} = 5.28 \text{ mN.m}^{-1}$). As for Triton X-100 and 350 351 Aerosol MA-80, the reductions were respectively 96% and 90% for a concentration of 64×CMC (σ_{final} 352 $= 0.48 \text{ mN.m}^{-1}$ and 1.10 mN.m $^{-1}$). Finally, SDBS seemed to be the most advantageous surfactant as it could reduce the IFT by 99% at a lower surfactant concentration ($\sigma_{\text{final}} = 0.10 \text{ mN.m}^{-1}$). Experimental 353 results showed that thermal enhancement had only limited influence on DNAPL/water IFT. This 354 parameter decreased by only 2.3% between 10 and 60 °C (respectively 12.17 to 11.89 mN.m⁻¹). 355

The DNAPL-water-glass contact angle without surfactant was 119.33 ± 4.16 ° at 20 °C (Supplementary materials, Figure A-3), which is typical of such a non-wetting fluid and close in order of magnitude to values reported in the literature (*e.g.*, 129° for TCE in pure water) [Orphius and Kibbey (2005)]. The reduction in contact angles with chemical enhancement were consistent with previous work [Amirpour et al. (2015)]. The contact angle was significantly reduced by adding SDBS; it reached 30.33° (*i.e.* a 75% drop) for SDBS concentrations far below the CMC and therefore DNAPL became wetting, as the angle was less than 90°. Regarding the other anionic surfactant, adding Aerosol MA-80 reduced the contact angle by 23% for a concentration equivalent to $16 \times \text{CMC}$ ($\theta_{\text{final}} = 91.33^{\circ}$). The two nonionic surfactants gave similar curves. Triton X-100 and Tween 80 only reduced the contact angles by 5.6 and 4.9% ($\theta_{\text{final}} = 112.70^{\circ}$ and 113.50°), respectively. The temperature increase only moderately reduced the contact angle. Increasing the temperature from 10 to 60 °C decreased the contact angle from 126.33 to 108.00° (*i.e.* a 14% reduction).

Based on these results, we can estimate the effect of chemical and thermal enhancements on capillary pressure, P_c. As a first approximation, we considered DNAPL and water in a tube (pore throat) with a circular cross section (Eq. 20).

$$P_c = \frac{2\sigma\cos\theta}{r}$$
 Eq. 20

371 where P_c (Pa) is the capillary pressure, $\sigma = \sigma_{nw}$ (mN.m⁻¹) is the interfacial tension, θ (°) is the pore 372 contact angle, and r (cm) is the mean radius of interface (r = r_{tube}/cos θ)

The nonionic surfactants reduced P_c less than the anionic surfactants. Triton X-100 and Tween 80 only reduced the capillary pressure by 77.3% and 62.5%, respectively, while the anionic surfactants were far more effective in reducing capillary pressure (respectively, 98.9% and 99.7% for Aerosol MA-80 and SDBS). Thermal enhancement had only a minor effect on P_c (49.4% reduction).

377

378 *4.1.3 Density*

The studied DNAPL was denser (1.66 kg.L⁻¹ at 20 °C) than well-known chlorinated solvents such as TCE and PCE (respective densities at 20 °C: 1.46 and 1.62 kg.L⁻¹). PCE density variations as a function of temperature are moderate; with decreases on the order of 0.1 kg.L⁻¹ for a temperature increase from 20 to 90 °C [Sleep and Ma (1997)]. The Tavaux site DNAPL also displayed a moderate density decrease as a function of temperature; as the density decreased only by 3% for temperatures ranging from 10 to 60 °C (Supplementary materials, Figure A-4).

385

386 *4.2 Solubilization*

The DNAPL solubility was low: 40.00 mg.L⁻¹ (at 20 °C) (Supplementary materials, Figure A-5). 387 388 The two non-ionic surfactants generated a global solubility increase (Σ [COC]) of a factor of 20 for Triton X-100 (for a concentration of 8×CMC_{theoretical}) whereas it was only 1.45 for Tween 80 for the 389 same concentration ratio. The addition of MA-80 aerosol at 8×CMC increased the solubilization 390 yielded by a factor of 18; but only by a factor of 1.3 for SDBS with the same CMC ratios. Usually, as 391 392 surfactant concentration increases, the volume of core micelles increases, leading to a linear increase in the apparent solubility [Pennell et al. (2014)]. A linear change in apparent solubility has also been 393 demonstrated for surfactant concentrations greater than their CMC for several chlorinated compounds 394 (e.g., PCE, TCE, HCBD and HCEa) [Jafvert (1994); Kommalapati et al. (1997); Harendra and 395 Vipulanandan (2011); Rodrigues et al. (2017)]. 396

The CMC_{real} of the surfactants that correspond to the IFT = f([surfactant]) and $[\Sigma COC]$ = f([surfactants]) curves' inflection points are (Supplementary materials, Figure A-5): 1200.00 mg.L⁻¹ for Triton X-100, 503.04 mg.L⁻¹ for Tween 80, 11640.00 mg.L⁻¹ for Aerosol MA-80 and 65.34 mg.L⁻¹ for SDBS. Increasing the temperature from 12 °C (temperature of the groundwater) to 60 °C increased the solubility by 13% (44.16 versus 50.00 mg.L⁻¹), which is relatively small change.

402

403 *4.3* Drainage-imbibition experiments

404 *4.3.1* Drainage-imbibition experiments without enhancement

The Pc-Sw curves for 0.5 and 0.1 mm GB are compared in Figure 2a. Table 1 shows the results of 405 these experiments and the van Genuchten-Mualem (VGM) fitting parameters. The experimental data 406 407 were fit to the VGM capillary pressure-saturation function (Eq. 1 and Eq. 2) using sum of squared 408 errors (SSE) minimization. The SSE were low (SSE<0.0014), which demonstrated that the VGM 409 model can be used to describe the experimental results. The experimental results are mainly focused 410 on residual DNAPL saturation and irreducible water saturation. It is clear from the graph that the parameters differed slightly for each size of glass bead. For 0.5 mm GB, S_m was approximately 10.9% 411 vs. 12.7% for 0.1 mm GB. According to the results, S_{rw} was around 24.8% for 0.5 mm GB and 30.9% 412 for 0.1 mm GB. This is explained by capillary forces as the mean pore radius was higher for 0.5 mm 413

GB and therefore the influence of capillary forces was lower. Therefore, during drainage, less DNAPL could be incorporated into the 0.5 mm GB medium (S_{rn} was 23.0% higher than for 0.1 mm GB) while during imbibition, more DNAPL trapped in the 0.5 mm GB medium could be extracted (S_{rn} was 16.0% higher than for 0.1 mm GB). The curve for 0.1 mm GB has a lower S_w amplitude. These results are in agreement with the data reported in the literature [Mualem (1976); van Genuchten (1980); Gerhard and Kueper (2003a)].

The parameter α value, characteristic of the reverse of the suction effects, was higher for the 0.5 mm GB, which was expected, since capillary effects were higher for the 0.1 mm GB. The slope of the median portion of the curves was relatively flat, which confirms that the GB were indeed homogeneous [Ouchiyama and Tanaka (1984); Likos and Jaafar (2013); Chapuis et al. (2015); Chiapponi (2017)]. The interpretation of parameter *n* values variations is more complex: sensitivity tests show that for high α values and *n* values greater than 4 (in a homogeneous medium), as in our case, the variations of n had a minor influence on the shape of the retention curves.

427

428 4.3.2 Drainage-imbibition experiments with enhancements

429 The remediation enhancement yield was calculated as follows (Eq. 21):

Remediation yield =
$$100 - 100 \left(\frac{S_{rn,chemical enhancement}}{S_{rn,without enhancement}} \right)$$
 Eq. 21

430 Adding SDBS had the highest impact on reducing the IFT. Therefore, the best remediation enhancement yield was obtained with SDBS (27.6% *i.e.* S_{rn} = 0.079), for 0.5 mm GB. Aerosol MA-80 431 and Triton X-100 had similar remediation enhancement yields (24.0 and 22.5% respectively, *i.e.* $S_m =$ 432 0.083 and 0.085). Tween 80 had a remediation enhancement yield close to zero ($S_m = 0.11$). For 0.1 433 434 mm GB, the best remediation enhancement yield was with SDBS (46.3% *i.e.* $S_{rn} = 0.068$). Aerosol MA-80 performed better than Triton X-100: the remediation enhancement yields were respectively 435 35.56 and 38.1% ($S_m = 0.082$ and 0.078). Tween 80 had a remediation enhancement yield of 7.0% (S_m 436 = 0.118). 437

The remediation enhancement yields for 0.1 mm GB were higher than those observed with the 0.5 mm GB. This is because the capillary forces were higher for 0.1 mm GB and the effect of the surfactants, whose purpose is to reduce the IFT, is improved. The best results were obtained with SDBS. Figure 2b shows the average of the experimental results for P_c -S_w curves with SDBS (imbibition) for the 0.5 mm GB and 0.1 mm GB (fitted with the VGM model). At the end of drainage, adding surfactant influences the slope of the line, which is slightly more horizontal, demonstrating that the capillary effects, and therefore the capillary fringes (thickness of the migration front), were weaker.

Capillary number (N_{ca}), Bond number (N_B), and Total trapping number (N_T) were calculated using 445 Eq. 7 to Eq. 9 on the basis of the experimental and rheological data (Figure 3). As expected, increasing 446 447 the grain size increases all three numbers. Here, viscosity forces predominate over gravitational forces, while capillary forces predominated over other forces. Adding surfactants increased N_{ca} and 448 consequently N_T. Increasing N_{ca} decreases the residual saturations [Lake (1989); Pennell et al. (1996); 449 Sheng (2015)]. It was found that the thermal enhancement had no effect on S_m or on the shape of the P_c-450 Sw curve during the imbibition stage of these static drainage-imbibition experiments. The thermal 451 enhancement only affected the viscosity. 452

453

454 4.4 Monitoring of the drainage-imbibition experiments with dielectric permittivity

455 4.4.1 Drainage-imbibition experiments without enhancement

Permittivity calibration curves were plotted as a function of water saturation for 0.5 and 0.1 mm GB (Figure 4). Permittivity could be related to water saturation by a second order polynomial relationship. The scatter plot was less dispersed for the 0.5 mm GB than for the 0.1 mm GB and therefore R^2 was lower for the 0.1 mm GB ($R^2 = 0.79$ vs 0.90 for 0.5 mm GB). This can be explained by the difference between capillary effects.

The resulting trends were very similar for 0.5 and for 0.1 mm GB. The calibration curves were also very similar: the ratios of polynomial curves for 0.5 mm GB/polynomial curves for 0.1 mm GB vary between 0.97 and 1.20. For homogeneous porous media, grain size has little influence on permittivity. These results were consistent with previous research on permittivities measured in a water-air system. The small differences can be attributed to the less regular GB, the spherical shape of GB and the pore connectivity [Robinson and Friedman (2001); Robinson and Friedman (2002); Robinson et al. (2005); Brovelli and Cassiani (2010)]. Experimental data was fitted to the CRIM model (see Eq. 12), using the following values: $\emptyset = 0.38$ (experimental values), $\varepsilon_w = 80$ (experimental value), $\varepsilon_n = 3.11$ (experimental value for 100% DNAPL), $\varepsilon_m = 7.5$ (literature review for GB materiel [von Hippel (1954); Robinson and Friedman (2002)].

We see that a value $\alpha_{\varepsilon} = 0.5$ (cited in the literature) does not fit our experiments. The α_{ε} values calculated by the least-square method were respectively 0.70 and 0.75 for 0.5 mm and 0.1 mm GB (with respective R² of 0.89 and 0.78). Therefore, for our experiments the CRIM model can be applied using the following equations (Eq. 22 to Eq. 23):

For 0.5 mm GB
$$\varepsilon' = \left[\emptyset \left(S_w \varepsilon'_w^{0.7} + S_n \varepsilon'_n^{0.7} \right) + (1 - \emptyset) \varepsilon'_m^{0.7} \right]^{\frac{1}{0.7}}$$
 Eq. 22

For 0.1 mm GB
$$\varepsilon' = \left[\phi \left(S_w \varepsilon'_w^{0.75} + S_n \varepsilon'_n^{0.75} \right) + (1 - \phi) \varepsilon'_m^{0.75} \right]^{\frac{1}{0.75}}$$
 Eq. 23

Figure 4 shows a satisfactory match between the experimental data and the CRIM model for higher saturations than S_{rw} (less than 8% difference). The CRIM model fitted the data better for 0.5 mm GB. For lower water saturation, the ratios between the values estimated and the experimental values increased. The estimated values were overestimated by a factor of 1.14 for 0.5 mm GB and were underestimated by a factor of 0.92 for 0.1 mm GB. Persson and Berndtsson (2002) also obtained a better match to their data by introducing an extra degree of freedom in the mixing equations [Persson and Berndtsson (2002)].

Figure 5 illustrates permittivity variations as a function of different drainage-imbibition cycles. The drainage-imbibition experiments could be closely correlated with permittivities but after the end of the second cycle (end of imbibition 2), the correlation no longer held. This is probably because the DNAPL and water ganglions that remain trapped during the experiments influenced the dielectric response.

Some authors have shown that the CRIM model based on the relative permittivity of free water, air (or non-wetting fluid) and solids may not be suitable in all situations, especially in the transition phases (when S_w are between S_{rn} and S_{rw}). These authors have proposed to consider a 4th constituent, the bound water. Indeed, the dielectric constant of bound water is very different from that of free water (due to the electrical bonds limiting the freedom of polarization of water molecules) [Dasberg and Hopmans (1992); Capparelli et al. (2018)]. An equation derived from the CRIM model incorporating the bound water has been proposed [Dobson et al. (1985); Dirksen and Dasberg (1993)]. Capparelli et al. (2018) carried out monitoring of S_w variations with TDR probes in unsaturated soils. They showed, as in our experimental results, a difference between the values estimated with the CRIM model and the experimental values. The experimental results were compared to the classical CRIM model, to the four-phase dielectric mixing model (with bound water), and to the classical CRIM model with a variable exponent. The best results were obtained with the three-phase dielectric mixing model with a variable exponent (α_e) [Capparelli et al. (2018)].

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- 501

4.4.2 Drainage-imbibition experiments with enhancements

502 Relative to the response considered as the blank test (water + GB), the mean measured permittivities were impacted by +1.4% (SDBS), +0.5% (Aerosol MA-80), -1.6% (Triton X-100), and -503 504 2.1% (Tween 80), respectively. It was observed, therefore, that the permittivity was slightly increased for nonionic surfactants and slightly decreased for anionic surfactants. But such deviations can be 505 506 considered as negligible. Figure 5 shows the averages of $S_{\rm m}$ with permittivity averages corresponding 507 to the experiments with and without surfactants (with 0.1 and 0.5 mm GB). Even though the standard deviations sometimes overlap, we can assess the recovery yields from the permittivity measurements: 508 the average permittivities for 0.5 and 0.1 mm GB without surfactants were respectively 27.77 and 509 510 25.32 (for S_w at the end of imbibition = 0.89 and 0.87). In presence of surfactants, the average 511 permittivities for 0.5 mm GB and 0.1 mm GB were respectively 29.07 and 30.81 (for S_w at the end of 512 imbibition of 0.915 and 0.917). Figure 5 also shows the permitivity variations that correspond to the 513 different remediation yields (for different surfactants): for the 0.5 mm GB, the final permittivities are, 514 in increasing order: $\rho_{c,SDBS} > \rho_{c,Aerosol MA-80} > \rho_{c,Triton X-100} > \rho_{c,Tween 80}$. The averages overestimations for 515 S_{rn} with chemical enhancement were by a factor of 1.06 ± 0.02 (for 0.5 mm GB) and 1.09 ± 0.01 (for 516 0.1 mm GB).

517 Experiments were conducted to estimate the influence of temperature on measured permittivities 518 for DNAPL and water with 0.5 mm GB (Supplementary materials, Figure A-6). The permittivity 519 values were influenced by temperature variations. Increasing the temperature from 20 to 50 °C caused 520 an average permittivity increase for DNAPL with GB of 19.0% (from 5.45 to 6.49). The variation in $ε_{DNAPL+GB}$ as a function of temperature, estimated from the slopes of the linear correlations, is + 0.034 m³.m⁻³.°C⁻¹. Results from various authors show that the permittivity of light chlorinated solvents fall slightly as temperature rises. A linear relation has been established, with coefficients ranging between -0.003 and -0.051 m³.m⁻³.°C⁻¹ [Morgan and Lowry (1930); Loon et al. (1967); Nath and Narain (1982); Nath (1995); Corradini et al. (1996); Ajo-Franklin et al. (2006)]. In our case, the DNAPL is essentially composed of a mixture of heavy chlorinated compounds, which may explain the difference in behavior.

Permittivities for water with GB decreased linearly as temperature rose (0.023 m³.m⁻³.°C⁻¹), with 528 values that were consistent with the literature [Weast (1986); Persson and Berndtsson (1998); Logsdon 529 (2000); Logsdon (2005); Seyfried and Grant (2007)]. The variations in $\varepsilon_{\text{DNAPL+GB}}$ and $\varepsilon_{\text{water+GB}}$ as a 530 function of temperature were quantified and can be interpreted with the CRIM model. At the end of 531 imbibition, the measured permittivities were 28.50 and 25.58 for 0.5 and 0.1 mm GB, respectively (vs 532 estimated permittivities of 26.89 and 27.12, respectively). The experiments with surfactants and 533 thermal enhancement showed that the reductions in residual saturations can be quantified with 534 535 permittivity measurements.

536

537 4.5 Monitoring of the drainage-imbibition experiments with electrical resistivity

538 4.5.1 Drainage-imbibition experiments without enhancement

539 Calibration curves ($\rho_c = f(S_w)$) were plotted for 0.5 and 0.1 mm GB (Figure 6). Experiments with DNAPL alone in GB were performed to determine the experimental values corresponding to $S_w = 0\%$ 540 and $S_w = 100\%$. Resistivity data were found to be more dispersed than permittivity data. Resistivity 541 542 can be correlated to water saturation via a power relationship. The correlation factor, considering all 543 the measured data, was satisfactory for 0.5 mm GB: $R^2 = 0.81$. The results were more dispersed with the 0.1 mm GB than with the 0.5 mm GB. Resistivity was related to S_w via a power law for 0.1 mm 544 GB; with a correlation factor $R^2 = 0.74$. The results were different for 0.5 and 0.1 mm GB: the features 545 of the trend curves are similar but are shifted. For homogeneous beads, and for identical Sw, the 546 547 resistivity was higher for the 0.1 mm GB. This can be explained by the difference in DNAPL distributions in pores. 548

549 On the basis of Eq. 14 to Eq. 17 and by applying the least-squares method, we obtained the 550 following equations (Eq. 24 and Eq. 25):

for 0.5 mm GB
$$\rho_{c,bulk} = \frac{1}{\sigma_{c,bulk}} = \frac{1}{\sigma_{c,DNAPL} [\emptyset(1-S_w)]^{1.752} + \sigma_{c,water} [\emptyset(S_w)]^{1.9}}$$
 Eq. 24

for 0.1 mm GB
$$\rho_{c,bulk} = \frac{1}{\sigma_{c,bulk}} = \frac{1}{\sigma_{c,DNAPL} [\emptyset(1-S_w)]^{1.678} + \sigma_{c,water} [\emptyset(S_w)]^{2.4}}$$
 Eq. 25

where $\rho_{c,bulk}$ (Ω .m) is the real effective electrical resistivity of the bulk, and $\rho_{c,DNAPL}$ (Ω .m) is the real effective electrical resistivity of DNAPL.

553 The conductivity values measured for DNAPL and water used to fit the model were: $\sigma_{c,DNAPL} = 3.1807 \times 10^{-8}$

554 S.m⁻¹ and $\sigma_{c.water} = 0.04$ S.m⁻¹.

555 A correction factor was used in order to better estimate the experimental data. The correction 556 factors were as follows (Eq. 26 and Eq. 27):

for 0.5 mm GB
$$\frac{\rho_{c,measured}}{\rho_{c,estimated}} = 1.34S_w^{0.40}$$
 Eq. 26

for 0.1 mm GB
$$\frac{\rho_{c,measured}}{\rho_{c,estimated}} = 2.00S_w - 0.011$$
 Eq. 27

557 Figure 7 illustrates the resistivity variations (measured and estimated) as a function of different 558 drainage-imbibition cycles. The model overestimated the values at the start and under-estimated the 559 values in most other cases. For the 0.5 mm GB, the average ratios of measured resistivity/estimated resistivity were 0.64 for $S_w = 100\%$ (start), 3.14 at the end of the drainage 1 and 0.99 at the end of the 560 imbibition 1. For the 0.1 mm GB, these ratios were 0.72 at the start, 12.84 at the end of drainage 1 and 561 562 5.22 at the end of imbibition 1. As for permittivity, it reminded that the resistivity measurement is very 563 local while the water saturation measurement is more global. Uncertainty is higher for large resistivity values because the measured potential difference is smaller. 564

These resistivity overestimations have been discussed by previous authors. Byun et al. (2019) showed that in a porous medium with water and gas, a cementation factor that considers the saturation condition improves resistivity modeling with Archie's law [Byun et al. (2019)]. Other authors also demonstrated that at pore scale, the resistivity indices did not generally obey Archie's law in the nonhomogeneous zones (for example in a transition zone with values close to S_{rn} and S_{rw}). It is therefore possible to over or underestimate resistivity depending on the thickness of the water films surrounding sand grains or GB [Li et al. (2015)]. These water films play a role not only on the pore space connectivity but also on resistivity [Bernabé et al. (2011); Li et al. (2015); Bernabé et al. (2016)]. Finally, previous authors have shown that many other parameters could be taken into account at different scales in electrical conductivity models in porous media: tortuosity, pore size distribution, pore-conductance distributions, interconnectivity, and universal power laws of percolation [Glover (2010); Cai et al. (2017); Ghanbarian and Sahimi (2017)].

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580 4.5.2 Drainage-imbibition experiments with enhancements

The experimental results show that adding surfactants generated a greater impact on resistivity 581 measurements than on permittivity measurements. For mean reference values of $132 \pm 17 \Omega$.m (for 582 water+GB), the mean values measured for nonionic surfactants were respectively $123 \pm 21 \Omega$.m for 583 584 SDBS and $30 \pm 9 \Omega$.m for Aerosol MA-80 (*i.e.* a factor of 0.92 and 0.22). Conversely, adding anionic surfactants increased resistivity: $216 \pm 25 \ \Omega$.m for Triton X-100 and $140 \pm 19 \ \Omega$.m for Tween 80, *i.e.* 585 respective increased of a factor of 1.62 and 1.05. However, these relative variations were small in 586 587 comparison with the resistivity of the DNAPL and GB ($\rho_{c,DNAPL} = 2730063$ and 3413632 Ω .m 588 respectively for 0.5 and 0.1 mm GB) and considering the high sensitivity and low accuracy of these geophysical measurements at very low frequencies. Such variations can therefore be considered as 589 590 negligible.

Figure 7 shows the averages of $S_{\rm m}$ with averaged resistivity corresponding to the experiments with and without surfactants (with the 0.1 and 0.5 mm GB). As shown, the residual saturations can be approached by the resistivity measurements. The correlations were less clear with resistivities than with permittivities. Decreasing trends for $S_{\rm m}$ are seen in the resistivity measurements. For the 0.5 mm GB, for example, the resistivities at the end of the imbibition with and without surfactant were respectively 150 and 84 Ω .m (for $S_{\rm m}$ of 0.109 and 0.085). For 0.1 mm GB, the differences were 597 greater: 1061 Ω .m (without surfactants with S_{rn} = 0.131) vs. 79 Ω .m (with surfactant with S_{rn} = 0.082). 598 The reported standard deviations were quite high for imbibition without surfactant.

599 Experiments were conducted to estimate the influence of temperature on measured resistivities of water and DNAPL. The resistivity of DNAPL at 50 °C only represented 25.3% of the resistivity at 20 600 °C (5.88×10⁵ vs. 2.32×10⁶ Ω .m, respectively at 50 and 20 °C). The resistivity values of water were 601 very low compared to those of DNAPL. Resistivity values were 31 Ω.m at 20 °C, and 17 Ω.m at 20 °C 602 603 (*i.e.* 57% of the value at 20 °C). The decrease in water resistivity as a function of temperature increase was due to the increase in ionic mobility [Dakhnov (1962); Hayashi (2004); Light et al. (2005); 604 Grellier et al. (2008)]. Dakhnov (1962) established the following relationship (for ionic fluids) (Eq. 605 606 28):

$$\rho_{c,w} = \frac{\rho_{c,w_0}}{1 + \alpha_{c,w}(T - T_0)}$$
 Eq. 28

607 where $\rho_{c,w}$ (Ω .m) is the electrical resistivity of the fluid at temperature *T* (°C), ρ_{c,w_0} (Ω .m) is the 608 resistivity of the fluid at temperature T₀, $\alpha_{c,w}$ is the temperature coefficient of resistivity ($\alpha_{c,w} \approx 0.023$ 609 °C⁻¹ for T₀ = 23 °C, and 0.025 °C⁻¹ for T₀ = 0 °C)

610 The slopes of the linear relationships in the graphs $(\rho_{c,0}/\rho_c)-1 = f(T-T_0)$ determine the values of the 611 coefficient α_c for the Dahhnov equation (Supplementary materials, Figure A-7).

With respect to the drainage-imbibition experiments with thermal enhancement, the residual saturations were almost identical with and without thermal enhancement. $\rho_{c,measured}$ values measured at the end of imbibition without enhancement were lower than during the thermal enhancement experiments: 63 Ω .m (vs. 150 Ω .m without thermal enhancement) for the 0.5 mm GB and 87 Ω .m (vs. 1061 Ω .m without thermal enhancement) for the 0.1 mm GB. At the end of imbibition 1, the $\rho_{c,estimated}$ were respectively 84 Ω .m for 0.5 mm GB (vs. 152 Ω .m at 20 °C) and 112 Ω .m for 0.1 mm GB (vs. 203 Ω .m at 20 °C). In theory, the $\rho_{c,measured}$ therefore decreased by 45% in both cases.

619

620 4.6 Monitoring of the drainage-imbibition experiments with image interpretation

Figure 8 shows an experiment with 0.5 mm GB. The pollutant was dark brown-black. Note that during the drainage, the migration front was relatively sharp (which corresponded to curve $Pc = f(S_w)$) for drainage 1 with a relatively horizontal plateau – Figure 2). At the end of drainage, the cell was filled with DNAPL but filling was not totally homogeneous (due to wall effects). During the imbibition, the migration front was more diffuse, which was expected considering the $P_c = f(S_w)$ curve for imbibition (which shows a less horizontal pseudo-plateau).

Figure A-8 (Supplementary materials) describes the variations of mean grey values for an 627 increasing AOI that begins at the centre of the picture, as detailed below. Increasing the AOI length 628 629 increased the number of contained pixels. This curve was calculated with a Fiji macro which served to associate a mean grey value with the S_w value, as the grey value was sensitive to the number of pixels 630 inside the selected AOI. However, for AOI between 500 pixels and 1000 pixels, the computed mean 631 grey values were stable, implying that between 2.5×10^5 and 1.0×10^6 pixels were necessary. This 632 interval length was chosen to compute all mean grey values for the calibration curve shown in Figure 633 9. 634

The calibration curve could only be made for the points corresponding to $S_w = 1, 0, S_m$ and S_{rw} . The 635 curve shows a good correlation ($R^2 = 0.98$). The size of the GB did not influence the calibration curve. 636 637 Also, no differences were observed in Optical Density (OD) measured for $S_w = 1$ and $S_w = 0$. S_m (with respective means of 0.24 and 0.32 for 0.5 and 0.1 mm GB) and S_{rw} values (with respective means of 638 0.89 and 0.84 for 0.5 and 0.1 mm GB) were very close to those measured with the circular 1D cells 639 640 (less than 3% difference). Variations in S_{rn} and S_{rw} follow a linear regression which is in agreement 641 with the literature [Schincariol et al. (1993); Flores et al. (2011); Watson et al. (2019)]. The optical density experiments show that residual saturations can be estimated accurately ($R^2 = 0.98$) even with 642 surfactants and thermal enhancement. The OD was found to vary linearly as a function of S_w. 643

644

645 5. Conclusions

Drainage-imbibition experiments were performed in 1D cells with the aim of comparing the efficiency of thermal and chemical enhancements on the DNAPL recovery performance. Four different surfactants (SDBS, Aerosol MA-80, Triton X-100 and Tween 80) were tested at concentrations well below their respective CMC (in order to avoid DNAPL dissolution). The best remediation enhancement yield was obtained with SDBS: 27.6% for the 0.5 mm GB and 46.3% for 0.1 mm GB. Experiments with thermal enhancement were also conducted at 50 °C (in order to avoid
DNAPL volatilization). However, no significant improvement in the remediation enhancement yield
was observed.

The drainage-imbibition experiments were monitored by electrical resistivity, permittivity and 654 optical density. The goal was to confirm whether the residual saturations could be estimated indirectly. 655 Regarding the permittivity measurements, the estimation of residual saturations fits well with the 656 657 CRIM model in most cases. The experiments with surfactants and thermal enhancement showed that 658 the variations in residual saturations could also be quantified with permittivity measurements. An adapted version of Archie's Law was used to model resistivity variations as a function of residual 659 660 saturation variations. The estimated electrical resistivity data showed less correlation with the measurements than the permittivity data. Also, it was not possible to accurately quantify water 661 662 saturations with electrical resistivity monitoring. However, the accuracy is sufficient (especially for high values of water saturation) for highlighting differences between the treatment technologies. The 663 optical density experiments showed that residual saturations can be accurately estimated even with 664 665 surfactant and thermal enhancement.

Future work will focus on multiphase flow modeling and pumping experiments in a 2 D tank at the laboratory scale, for the purpose of process up-scaling. The permittivity and resistivity measurements as well as the optical densities during the pumping test over time shall provide information regarding how chemical enhancement affects DNAPL recovery yields (at different flow rates). At field scale, the combined monitoring of electrical resistivities (which provide a global picture) and permittivities (which provide accurate but spatially limited information) is expected to improve the reliability of residual saturation data.

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- 689 [ADEME and Ernst & Young (2014)] ADEME and Ernst & Young: 2014, *Taux d'utilisation et coûts*690 *des différentes techniques et filières de traitement des sols et des eaux souterraines polluées en France*
- 691 Synthèse des données 2012, ADEME edn, Agence de l'Environnement et de la Maîtrise de
 692 l'Energie, Agence de l'Environnement et de la Maîtrise de l'Energie 20, avenue du Grésillé BP
- 693 90406 49004 Angers Cedex 01. 148 p.
- [Ahn et al. (2008)] Ahn, C., Kim, Y., Woo, S. and Park, J.: 2008, Soil washing using various
 nonionic surfactants and their recovery by selective adsorption with activated carbon, *Journal of Hazardous Materials* 154(1–3), 153–160.
- [Ajo-Franklin et al. (2004)] Ajo-Franklin, J., Geller, J. and Harris, J.: 2004, The dielectric
 properties of granular media saturated with DNAPL/water mixtures, *Geophysical Research Letters* **31**(L17501), 1–4.
- [Ajo-Franklin et al. (2006)] Ajo-Franklin, J., Geller, J. and Harris, J.: 2006, A survey of the
 geophysical properties of chlorinated DNAPLs, *Journal of Applied Geophysics* 59, 177–189.
- [Alazaiza et al. (2016)] Alazaiza, M., Nigien, S., Ishak, W. and Kamaruddin, S.: 2016, A review of
 light reflection and transmission methods in monitoring non-aqueous phase liquid migration in porous
 media, *Journal of Engeneering and Applied Sciences* 11(4), 2319–2326.
- [Amirpour et al. (2015)] Amirpour, M., Shadizadeh, S., Esfandyari, H. and Ahmadi, S.: 2015,
 Experimental investigation of wettability alteration on residual oil saturation using nonionic
 surfactants: Capillary pressure measurement, *Petroleum* 1(4), 289–299.
- [Andersson et al. (2014)] Andersson, M., Bennetzen, M., Klamt, A. and Stipp, S.: 2014, FirstPrinciples Prediction of Liquid/Liquid Interfacial Tension, *Journal of Chemical Theory and Computation* 10, 3401–3408.
- [Archie (1942)] Archie, G.: 1942, The electrical resistivity log as an aid in determining some reservoir
 characteristics, *Petroleum Transactions of AIME* 146(1), 54–62.
- 713 [Atteia et al. (2013)] Atteia, O., Del Campo Estrada, E. and Bertin, H.: 2013, Soil flushing: a
 714 review of the origin of efficiency variability, *Reviews in Environmental Science and Bio/Technology*715 12, 379–389.

- 716 [Berglund and Cvetkovic (1995)] Berglund, S. and Cvetkovic, V.: 1995, Pump-and-Treat
 717 Remediation of Heterogeneous Aquifers: Effects of Rate-Limited Mass Transfer, *Groundwater*718 33, 675–685.
- [Bernabé et al. (2016)] Bernabé, Y., Li, M., Tang, Y. and Evans, B.: 2016, Pore space connectivity
 and the transport properties of rocks, *Oil & Gas Science and Technology* **71**(4), 1–17.
- [Bernabé et al. (2011)] Bernabé, Y., Zamora, M., Li, M., Maineult, A. and Tang, Y.: 2011, Pore
 connectivity, permeability, and electrical formation factor: A new model and comparison to
 experimental data, *Journal of Geophysical Research* 116(B11204), 1–15.
- [Birchak et al. (1974)] Birchak, J., Gardner, C., Hipp, J. and Victor, J.: 1974, High dielectric constant
 microwave probes for sensing soil moisture, *Proceedings of the Institute of Electrical and Electronics Engineers* 62(1), 93–98.
- 727 [Brewster et al. (1995)] Brewster, M., Annan, A., Greenhouse, J., Kueper, B., Olhoeft, G., Redman, J.
- and Sander, K.: 1995, Observed migration of a controlled DNAPL release by geophysical methods, *Ground Water* 33, 987–997.
- [Brovelli and Cassiani (2010)] Brovelli, A. and Cassiani, G.: 2010, A combination of the hashinshtrikman bounds aimed at modelling electrical conductivity and permittivity of variably saturated
 porous media, *Geophysical Journal International* 180, 225–237.
- [Byun et al. (2019)] Byun, Y., Hong, W. and Yoon, H.: 2019, Characterization of cementation
 factor of unconsolidated granular materials through time domain reflectometry with variable saturated
 conditions, *Materials* 12(8), 1340–1354.
- [Cai et al. (2017)] Cai, J., Wei, W., Hu, X. and Wood, D.: 2017, Electrical conductivity models
 in saturated porous media: A review, *Earth-Science Reviews* 171, 419–433.
- [Capparelli et al. (2018)] Capparelli, G., Spolverino, G. and Greco, R.: 2018, Experimental
 Determination of TDR Calibration Relationship for Pyroclastic Ashes of Campania (Italy), *Sensors*18(3727), 1–14.
- 741 [Cazaux et al. (2014)] Cazaux, D., Colombano, S., Joubert, A., Dumestre, A. and Lecuelle, G.: 2014,
- 742 Optimized physical recovery of DNAPL using upwelling technique and geostatistical analysis at large

- field scale, *Ninth International Conference on Remediation of Chlorinated and Recaclitrant Compounds*, number A1, Battelle Press, Columbus, OH, USA, Monterey, CA, USA, p. 5.
- [Chambers et al. (2004)] Chambers, J., Loke, M., Ogilvy, R. and Meldrum, P.: 2004,
 Noninvasive monitoring of DNAPL migration through a saturated porous medium using electrical
 impedance tomography, *Journal of Contaminant Hydrology* 68(1–2), 1–22.
- 748 [Chapuis et al. (2015)] Chapuis, R., Weber, S. and Duhaime, F.: 2015, Permeability test results with
- packed spheres and non-plastic soils, *Geotechnical Testing Journal* **38**(6), 950–964.
- 750 [Chiapponi (2017)] Chiapponi, L.: 2017, Water retention curves of multicomponent mixtures of
 751 spherical particles, *Powder Technology* 320, 646–655.
- 752 [Childs et al. (2004)] Childs, J., Acosta, E., Knox, R., Harwell, J. and Sabatini, D.: 2004, Improving
- the extraction of tetrachloroethylene from soil columns using surfactant gradient systems, Journal of
- 754 *Contaminant Hydrology* **71**(1-4), 27–45.
- 755 [Cohen and Mercer (1993)] Cohen, R. and Mercer, J.: 1993, DNAPL Site Evaluation, *Technical*756 *Report EPA/600/R-93/022*, Edited by USEPA Office of Research and Development. 369 p.
- [Constable and Srnka (2007)] Constable, S. and Srnka, L.: 2007, An introduction to marine
 controlled-source electromagnetic methods for hydrocarbon exploration, *Geophysics* 72(2), WA3–
 WA12.
- [Corradini et al. (1996)]Corradini, F., Marchetti, A., Tagliazucchi, M. and Tassi, L.: 1996, Static
 dielectric constants of 1,2-dichloroethane + 2-methoxyethanol + 1,2-dimethoxyethane ternary liquid
 mixtures from 10 to 80 °C, *Fluid Phase Equilibria* 124(1-2), 209–220.
- 763 [Dahlin (2000)] Dahlin, T.: 2000, Short note on electrode charge-up effects in dc resistivity data
 764 acquisition using multielectrode arrays, *Geophysical Prospecting* 48(1), 181–187.
- 765 [Dakhnov (1962)] Dakhnov, V.: 1962, *Geophysical well logging*, number 57-2, Q. Colorado
 766 School of Mines. 445 p.
- 767 [Dasberg and Hopmans (1992)] Dasberg, S. and Hopmans, J.: 1992, Time domain reflectometry
 768 calibration for uniformly and nonuniformly wetted sandy and clayey loam soils, *Soil Science Society*769 *of America Journal* 56(5), 1341–1345.

[Davis (1997)] Davis, E.: 1997, How Heat Can Enhance In-situ Soil and Aquifer
Remediation:Important Chemical Properties and Guidance on Choosing the Appropriate Technique, *Technical Report EPA/540/S-97/502*, U.S. Environmental Protection Agency, U.S. Environmental
Protection Agency; Solid Waste and Emergency Response, Washington, DC, USA. 18 p.

[Deparis et al. (2019)] Deparis, J., Joubert, A., Francois, B., Nodot, E., Invernizzi, T., Iravani, A.,

775 Dumestre, A., Fatin-Rouge, N., Maire, J., Kaifas, D., Triger, A., Klein, P., Giraud, Q., Paris, B.,

Cazaux, D., Gourry, J., Davarzani, H. and Colombano, S.: 2019, On the use of Geophysical
measurements to monitored DNAPL extraction, *AquaConSoil 2019, 15th International Conference*,
Antwerp, Belgium.

[Dirksen and Dasberg (1993)] Dirksen, C. and Dasberg, S.: 1993, Improved calibration of time
domain reflectometry soil water content measurements, *Soil Science Society of America Journal*57(3), 660–667.

[Dobson et al. (1985)] Dobson, M., Ulaby, F., Hallikainen, M. and El-Rayes, M.: 1985, Microwave
dielectric behaviour of wet soil, part ii, dielectric mixing models, *Institute of Electrical and Electronics Engineers Transactions on Geoscience and Remote Sensing* 23, 35–46.

[Duffield and Ramamurthy (2003)] Duffield, A. and Ramamurthy, R.: 2003, Surfactant Enhanced
Mobilization of Mineral Oil within Porous Media, *Water, Air, and Soil Pollution* 143(1–4), 111–122.

787 [Dullien (1992)]Dullien, F.: 1992, Porous Media: Fluid Transport and Pore Structure, number

788 978-0-12-223651-8, 2nd edn, Academic Press, San Diego, CA, USA. 574 p.

789 [Dwarakanath et al. (1999)] Dwarakanath, V., Kostarelos, K., Pope, G., Shotts, G. and Wade, W.:

790 1999, Anionic surfactant remediation of soil columns contaminated by nonaqueous phase liquids,

791 *Journal of Contaminant Hydrology* **38**(4), 465–488.

792[Endres and Knight (1992)]Endres, A. and Knight, R.: 1992, A theoretical treatment of the effect

of microscopic fluid distribution on the dielectric properties of partially saturated rocks, *Geophysical Prospecting* 40(3), 307–324.

795 [Falta et al. (2005b)] Falta, R., Basu, N. and Rao, P.: 2005b, Assessing impacts of partial mass

depletion in DNAPL source zones: II. Coupling source strength functions to plume evolution, *Journal*

797 *of Contaminant Hydrology* **79**(1-2), 45–66.

- [Falta et al. (2005a)] Falta, R., Rao, P. and Basu, N.: 2005a, Assessing the impacts of partial mass
 depletion in DNAPL source zones: I. Analytical modeling of source strength functions and plume
 response, *Journal of Contaminant Hydrology* 78(4.), 259–280.
- [Flores et al. (2011)] Flores, G., Katsumi, T., Inui, T. and Kamon, M.: 2011, A simplified image
 analysis method to study LNAPL migration in porous media, soils and foundations, *Soils and Foundations* 51(5), 35–847.
- 804 [Gerhard and Kueper (2003a)] Gerhard, J. and Kueper, B.: 2003a, Capillary pressure characteristics
- necessary for simulating DNAPL infiltration, redistribution, and immobilization in saturated porous
 media, *Water Resources Research* 39(8), 1–17.
- 807 [Ghanbarian and Sahimi (2017)] Ghanbarian, B. and Sahimi, M.: 2017, Electrical conductivity
 808 of partially saturated packings of particles, *Transport in Porous Media* 118(1), 1–16.
- [Glover (2010)]Glover, P.: 2010, A generalised archie's law for n phases, *Geophysics* 6, E247–E265.
- [Grellier et al. (2008)] Grellier, S., Guérin, R., Robain, H., Bobachev, A., Vermeersch, F. and
 Tabbagh, A.: 2008, Monitoring of leachate recirculation in a bioreactor landfill by 2-d electrical
 resistivity imaging, *Journal of Environmental and Engineering Geophysics* 13(4), 351–359.
- [Han et al. (2015)] Han, T., Best, A., Sothcott, J., North, L. and MacGregor, L.: 2015,
 Relationships among low frequency (2 hz) electrical resistivity, porosity, clay content and
 permeability in reservoir sandstones, *Journal of Applied Geophysics* 112, 279–289.
- 816 [Harendra and Vipulanandan (2011)] Harendra, S. and Vipulanandan, C.: 2011, Solubilization and
 817 degradation of perchloroethylene (PCE) in cationic and nonionic surfactant solutions, *Journal of*
- 818 Environmental Sciences **23**(8), 1240–1248.
- 819 [Harkness and Konzuk (2014)] Harkness, M. and Konzuk, J.: 2014, Cost analyses for remedial
- 820 options, in Chapter 16 in Chlorinated Solvent Source Zone Remediation, number 978-1-4614-6921-6,
- 821 Springer, SERDP ESTCP Environmental Remediation Technology, New York, NY, USA. 713 p.
- 822 [Hayashi (2004)] Hayashi, M.: 2004, Temperature-electrical conductivity relation of water for
- 823 environmental monitoring and geophysical data inversion, *Environmental Monitoring and Assessment*
- **96**(1-3), 119–128.

- 825 [Huang et al. (2015)] Huang, J., Christ, J., Goltz, M. and Demond, A.: 2015, Modeling NAPL
- dissolution from pendular rings in idealized porous media, *Water Resources Research* **51**, 8182–8197.
- 827 [IARC (2018)] IARC: 2018. https://www.iarc.fr/
- 828 [ITRC (2002)] ITRC: 2002, DNAPL Source Reduction: Facing the Challenge, Technical/Regulatory
- 829 Guidelines, Interstate Technology & Regulatory Council, Washington, DC, USA. 40 p.
- 830 [Jafvert (1994)] Jafvert, C.: 1994, Solubilization of non-polar compounds by non-ionic surfactant
- 831 micelles, *Water Research* **28**(5), 1009–1017.
- [Kargas and Soulis (2012)] Kargas, G. and Soulis, K.: 2012, Performance analysis and calibration
 of a new low-cost capacitance soil moisture sensor, *Journal of Irrigation and Drainage Engineering*138(7), 632–641.
- 835 [Kommalapati et al. (1997)] Kommalapati, R., Valsaraj, K., Constant, W. and Roy, D.: 1997,
- 836 Aqueous solubility enhancement and desorption of hexachlorobenzene from soil using a plant-based
- 837 surfactant, *Water Research* **31**(9), 2161–2170.
- 838 [Kong (2004)] Kong, L.: 2004, Characterization of mineral oil, coal tar and soil properties and
- 839 investigation of mechanisms that affect coal tar entrapment in and removal from porous media, PhD
- 840 thesis, School of Civil and Environmental Engineering, Georgia Institute of Technology, Atlanta,
- 841 Atlanta, GA, USA. 309 p. https://smartech.gatech.edu/handle/1853/5093?show=full
- 842 [Kueper et al. (1993)] Kueper, B., Redman, D., Starr, R., Reitsma, S. and Mah, M.: 1993, A field
- 843 experiment to study the behavior of tetrachlorethylene below the water table: Spatial distribution of
- residual and pooled DNAPL, *Ground Water* **31**, 756–766.
- [Kueper et al. (2003)] Kueper, B., Wealthall, G., Smith, J., Leharne, S. and Lerner, D.: 2003, An
- 846 illustrated handbook of DNAPL transport and fate in the subsurface, number 1844320669, U.K.
- 847 Environment Agency, Bristol, England, United Kingdom. 67 p.
- 848 [Lake (1989)] Lake, L.: 1989, Enhanced Oil Recovery, number 0132816016 9780132816014,
- 849 Prentice Hall, Old Tappan, NJ, USA. 550 p.
- 850 [Lenormand et al. (1988)] Lenormand, R., Touboul, E. and Zarcone, C.: 1988, Numerical models
- and experiments on immiscible displacements in porous media, Journal of Fluid Mechanics 189, 165–
- 852 187.

- [Li et al. (2015)] Li, M., Tang, Y., Bernabé, Y., Zhao, J., Li, X., Bai, X. and Zhang, L.: 2015,
 Pore connectivity, electrical conductivity, and partial water saturation: Network simulations, *Journal of Geophysical Research: Solid Earth* 120(6), 4055–4068.
- [Light et al. (2005)] Light, T., Licht, S., Bevilacqua, A. and Morashc, K.: 2005, The fundamental
 conductivity and resistivity of water, *Electrochemical and Solid-State Letters* 8(1), E16–E19.
- 858 [Likos and Jaafar (2013)] Likos, W. and Jaafar, R.: 2013, Pore-scale model for water retention
- and fluid partitioning of partially saturated granular soil, *Journal of Geotechnical and Geoenvironmental Engineering* 139(5), 724–737.
- 861 [Logsdon (2000)] Logsdon, S.: 2000, Effect of cable length on time domain reflectometry
 862 calibration for high surface area soils, *Soil Science Society of America Journal* 64(1), 54–61.
- 863 [Logsdon (2005)] Logsdon, S.: 2005, Soil dielectric spectra from vector network analyzer data,
 864 Soil Science Society of America Journal 69(4), 983–989.
- [Loon et al. (1967)] Loon, R., Fuks, S. and Bellemans, A.: 1967, Dielectric constant of carbon
 tetrachloride and 1,1,1-trichloroethane mixtures, *Bulletin des Societes Chimiques Belges* 76(3-4), 202–
 210.
- [Luciano et al. (2010)] Luciano, A., Viotti, P. and Papini, M. P.: 2010, Laboratory investigation of
 DNAPL migration in porous media, *Journal of Hazardous Materials* 176(1-3), 1006–1017.
- 870 [Mackay and Cherry (1989)] Mackay, D. and Cherry, J.: 1989, Groundwater contamination: pump-
- and-treat remediation, *Environmental Science & Technology* **23**, 630–636.
- [Maire et al. (2018)] Maire, J., Joubert, A., Kaifas, D., Invernizzi, T., Mardue, J., Colombano, S.,
- 873 Cazaux, D., Marion, C., Klein, P., Dumestre, A. and Fatin-Rouge, N.: 2018, Assessment of flushing
- 874 methods for the removal of heavy chlorinated compounds DNAPL in an alluvial aquifer, *Science of*
- 875 *the Total Environment* (612), 1149–1158.
- [McDade et al. (2005)] McDade, J., McGuire, T. and Newell, C.: 2005, Analysis of DNAPL source-
- depletion costs at 36 field sites, *Remediation Journal* **15**(2:), 9–18.
- 878 [McGuire et al. (2006)] McGuire, T., McDade, J. and Newell, C.: 2006, Performance of DNAPL
- 879 Source Depletion Technologies at 59 Chlorinated Solvent-Impacted Sites, Ground Water Monitoring
- and Remediation **26**(1), 73–84.

- [Morgan and Lowry (1930)] Morgan, S. and Lowry, H.: 1930, Dielectric polarization of some pure
 organic compounds in the dissolved, liquid, and solid states, *The Journal of Physical Chemistry*34(11), 2385–2432.
- [Mualem (1976)] Mualem, Y.: 1976, A new model for predicting the hydraulic conductivity of
 unsaturated porous media, *Water Resources Research* 12(3), 513–522.
- [Nath (1995)] Nath, J.: 1995, Ultrasonic velocities, relative permittivities, and refractive indices for
- binary liquid mixtures of trichloroethylene with pyridine and quinoline, *Fluid Phase Equilibria*109(1), 39–51.
- [Nath and Narain (1982)] Nath, J. and Narain, B.: 1982, Binary systems of tetrachlorethylene
 with benzene, toluence, p-xylene, carbon tetrachloride, and cyclohexane. 1. Ultrasonic velocities and
 adiabatic compressibilities at 293.15 and 303.15 K, dielectric constants at 298.15 and 308.15 K, and
 refractive indexes at 298.15 K, *Journal of Chemical Engineering Data* 27(3), 308–312.
- [Newell and Adamson (2005)] Newell, C. and Adamson, D.: 2005, Planning-level source decay
 models to evaluate impact of source depletion on remediation timeframe, *Remediation* 15(4), 27–47.
- 895 [NIEHS (2015)] NIEHS: 2015, Chlorinated organics information page, website.
 896 http://tools.niehs.nih.gov/srp/research/research4_s3_s4.cfm
- [Noel (2014)] Noel, C.: 2014, Suivi de la biodégradation des hydrocarbures par le couplage des
 mesures géophysiques électriques du sol (polarisation provoquée) et des analyses des gaz
 (concentration du CO2 et isotopie du carbone), PhD thesis, Université d'Orléans, Orléans, France.
 256 p.
- 901 [O'Carroll et al. (2004)] O'Carroll, D., Bradford, S. and Abriola, L.: 2004, Infiltration of PCE
 902 in a system containing spatial wettability variations, *Journal of Contaminant Hydrology* 73(1-4), 39–
 903 63.
- 904 [Orphius and Kibbey (2005)] Orphius, I. and Kibbey, T.: 2005, Dissolution-Induced Contact Angle
 905 Modification in Dense Nonaqueous Phase Liquid (DNAPL)/Water Systems, *Environmental Science*906 *and Technology* **39**, 1698–1706.

- 907 [Ouchiyama and Tanaka (1984)] Ouchiyama, N. and Tanaka, T.: 1984, Porosity estimation for
 908 random packings of spherical particles, *Industrial & Engineering Chemistry Fundamentals* 23, 490–
 909 493.
- 910 [Pankow and Cherry (1996)] Pankow, J. and Cherry, J.: 1996, Dense Chlorinated Solvents and
 911 Other DNAPLs in Groundwater: History, Behavior, and Remediation, number 978-0964801417,
- 912 Waterloo Press, Portland, OR, USA. 525 p.
- 913 [Pennell et al. (2014)] Pennell, K., Capiro, N. and Walker, D.: 2014, Surfactant and cosolvent
 914 flushing, Chapter 11 in Chlorinated Solvent Source Zone Remediation, number 978-1-4614-6921-6,
- 915 Springer, SERDP ESTCP Environmental Remediation Technology, New York, NY, USA. 713 p.
- 916 [Pennell et al. (1996)] Pennell, K., Pope, G. and Abriola, L.: 1996, Influence of viscous and
 917 buoyancy forces on the mobilization of residual tetrachloroethylene during surfactant flushing,
 918 *Environmental Science & Technology* 30, 1328–1335.
- 919 [Persson and Berndtsson (1998)] Persson, M. and Berndtsson, R.: 1998, Texture and electrical
 920 conductivity effects on temperature dependency in time domain reflectometry, *Soil Science Society of*921 *America Journal* 62(4), 887–893.
- 922 [Persson and Berndtsson (2002)] Persson, M. and Berndtsson, R.: 2002, Measuring nonaqueous
 923 phase liquid saturation in soil using time domain reflectometry, *Water Resources Research*924 38(5), 22.1–22.8.
- 925 [Redman and DeRyck (1994)] Redman, J. and DeRyck, S.: 1994, Monitoring non-aqueous phase
 926 liquids in the subsurface with multilevel time domain reflectometry probes, *Proceedings of the*927 Symposium on Time Domain Reflectometry in Environmental, Infrastructure, and Mining
 928 Applications, Spec. Publ. SP, .NTIS PB95-105789, 19-94, U.S. Bur. of Mines, Washington, D.C.,
 929 USA, p. 207–215.
- [Redman et al. (1991)] Redman, J., Kueper, B. and Annan, A.: 1991, Dielectric stratigraphy of a
 DNAPL spill and implications for detection with ground penetrating radar, *Aquifer Restoration*, *Ground Water Monitoring and Geophysical Methods*, *5th National Outdoor Action Conference*, Natl.
- 933 Ground Water Assoc., Las Vegas, NV., USA.

- 934 [Revil (2012)] Revil, A.: 2012, Spectral induced polarization of shaly sands: influence of the
 935 electrical double layer, *Water Resources Research* 48(2), 1–23.
- 936 [Reynolds (2011)] Reynolds, J.: 2011, An Introduction to Applied and Environmental
 937 *Geophysics*, number 978-0-471-48535-3, 2nd edition edn, Wiley-Blackwell, Chichester, UK. 710 p.
- [Robinson and Friedman (2001)] Robinson, D. and Friedman, S.: 2001, Effect of particle size
 distribution on the effective dielectric permittivity of saturated granular media, *Water Resources Research* 37(1), 33–40.
- [Robinson and Friedman (2002)] Robinson, D. and Friedman, S.: 2002, The effective
 permittivity of dense packings of glass beads, quartz sand and their mixtures immersed in different
 dielectric backgrounds, *Journal of Non-Crystalline Solids* 305, 261–267.
- [Robinson et al. (2005)]Robinson, D., Jones, S., Blonquist, J. and Friedman, S.: 2005, A physically
 derived water content/permittivity calibration model for coarse-textured, layered soils, *Soil Science Society of America Journal* 69(5), 1372.
- 947 [Rodrigues et al. (2017)] Rodrigues, R., Betelu, S., Colombano, S., Masselot, G., Tzedakis, T.
 948 and Ignatiadis, I.: 2017, Influence of temperature and surfactants on the solubilization of
 949 hexachlorobutadiene and hexachloroethane, *Journal of Chemical & Engineering Data* 62(10), 3252–
 950 3260.
- [Rosen (1989)] Rosen, M.: 1989, *Surfactants and Interfacial Phenomena*, 2nd edn, John Wiley and
 Sons, New York, NY, USA. 431 p.
- [Roth et al. (1990)] Roth, K., Schulin, R., Fluhler, H. and Attinger, W.: 1990, Calibration of time
 domain reflectometry for water content measurements using a composite dielectric approach, *Water Resources Research* 26, 2267–2273.
- [Sabatini et al. (1998)] Sabatini, D., Harwell, J., Hasegawa, M. and Knox, R.: 1998, Membrane
 processes and surfactant-enhanced subsurface remediation: results of a field demonstration, *Journal of Membrane Science* 151(1), 87–98.
- 959 [Sabatini et al. (2000)] Sabatini, D., Knox, R., Harwell, J. and Wu, B.: 2000, Integrated design of
- 960 surfactant enhanced DNAPL remediation: Efficient supersolubilization and gradient systems, *Journal*
- 961 *of Contaminant Hydrology* **45**(1-2), 99–121.

- 962 [Sale (2001)] Sale, T.: 2001, Methods for Determining Inputs to Environmental Petroleum
 963 Hydrocarbon Mobility and Recovery Models, *Technical Report API PUBLICATION 4711*, American
 964 Petroleum Institute, 1220 L Street, Northwest., Washington, DC, USA. 72 p.
- 965 [Schincariol et al. (1993)] Schincariol, R., Herderick, E. and Schwartz, F.: 1993, On the 966 application of image analysis to determine concentration distributions in laboratory experiments,
- 967 *Journal of Contaminant Hydrology* **12**(3), 197–215.
- 968 [Schincariol and Schwartz (1990)] Schincariol, R. and Schwartz, F.: 1990, An experimental
 969 investigation of variable density flow and mixing in homogeneous and heterogeneous media, *Water*970 *Resources Research* 26(10), 2317–2329.
- 971 [Schwille (1988)] Schwille, F.: 1988, Dense Chlorinated Solvents in Porous and Fractured
- 972 *Media Model Experiments*, number 978-0873711210, Lewis Publishers, Chelsea, MI, USA. 146 p.
- 973 [Seyfried and Grant (2007)] Seyfried, M. and Grant, L.: 2007, Temperature effects on soil
 974 dielectric properties measured at 50 mhz, *Vadose Zone Journal* 6(4), 759–765.
- 975 [She and Sleep (1998)] She, H. and Sleep, B.: 1998, The effect of temperature on capillary pressure976 saturation relationships for air-water and perchloroethylene-water systems, *Water Resources Research*
- **977 34**(10), 2587–2597.
- 978 [Sheng (2015)] Sheng, J.: 2015, Status of surfactant EOR technology, *Petroleum* 1, 97–105.
- [Sleep and Ma (1997)] Sleep, B. and Ma, Y.: 1997, Thermal variation of organic fluid properties and
 impact on thermal remediation feasibility, *Journal of Soil Contamination* 6(3), 281–306.
- 981 [Stimson (1974)] Stimson, A.: 1974, Photometry and Radiometry for Engineers, number 978-
- 982 0471825319, John Wiley & Sons Inc, New York, NY, USA. 488 p.
- 983 [Stroo et al. (2012)] Stroo, H., Leeson, A., Marqusee, J., Johnson, P., Ward, C., Kavanaugh, M.,
- Sale, T., Newell, C., Pennell, K., Lebrón, C. and Unger, M.: 2012, Chlorinated ethene source
 remediation: lessons learned, *Environmental Science & Technology* 46(6438–6447).
- [Stroo et al. (2003)] Stroo, H., Unger, M., Ward, C., Kavanaugh, M., Vogel, C., Leeson, A.,
 Marqusee, J. and Smith, B.: 2003, Peer Reviewed: Remediating Chlorinated Solvent Source Zones, A
 workshop lists the challenges and research needs, *Environmental Science & Technology*37(11), 224A–230A.

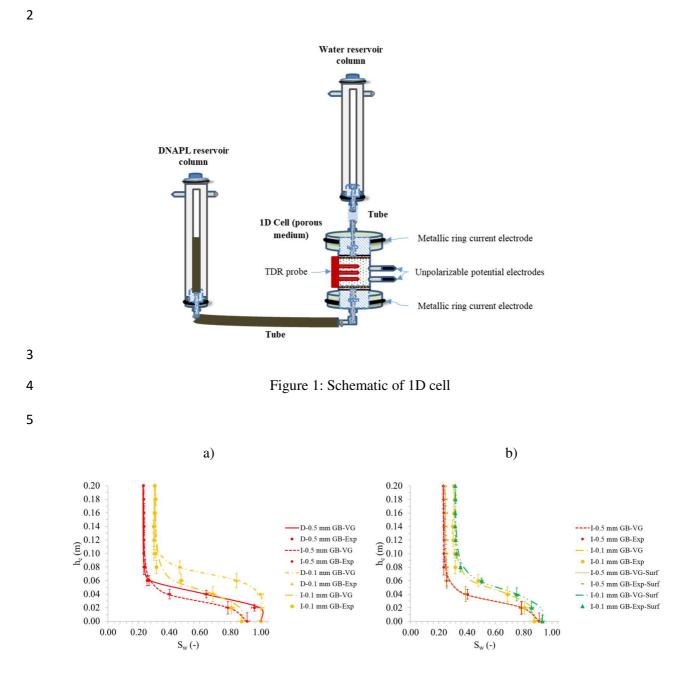
- 990 [Stupp and Paus (1999)] Stupp, H. and Paus, L.: 1999, Migrationsverhalten organischer
 991 Grundwasser-Inhaltsstoffe und daraus resultierende Ansätze zur Beurteilung von Monitored Natural
 992 Attenuation (MNA), *TerraTech* (5).
- [Suchomel et al. (2007)]
 Suchomel, E., Ramsburg, C. and Pennell, K.: 2007, Evaluation of
 trichloroethene recovery processes in heterogeneous aquifer cells flushed with biodegradable
 surfactants, *Journal of Contaminant Hydrology* 94(3-4), 195–214.
- [Taylor et al. (2001)] Taylor, T., Pennell, K., Abriola, L. and Dane, J.: 2001, Surfactant enhanced
 recovery of tetrachloroethylene from a porous medium containing low permeability lenses: 1.
 Experimental studies, *Journal of Contaminant Hydrology* 48(3–4), 325–350.
- 999 [Travis and Doty (1990)] Travis, C. and Doty, C.: 1990, Can contaminated aquifers at
 1000 Superfund sites be remediated?, *Environmental Science & Technology* 24, 1464–1466.
- 1001 [van Genuchten (1980)]van Genuchten, M.: 1980, A closed-form equation for predicting the hydraulic
- 1002 conductivity of unsaturated soils, *Soil Science Society of America* **44**, 892–898.
- 1003 [von Hippel (1954)] von Hippel, A.: 1954, *Dielectric Materials and Applications*, Technology
 1004 Press of M.I.T. and Wiley, Hoboken, N J, USA. 456 p.
- 1005 [Watson et al. (2019)] Watson, F., Maes, J., Geiger, S., Mackay, E., Singleton, M., McGravie, T.,
- 1006 Anouilh, T., Jobe, T., Zhang, S., Agar, S., Ishutov, S. and Hasiuk, F.: 2019, Comparison of Flow and
- 1007 Transport Experiments on 3D Printed Micromodels with Direct Numerical Simulations, Transport in
- 1008 *Porous Media* **129**(2), 449–466.
- 1009 [Weast (1986)] Weast, R.: 1986, CRC handbook of chemistry and physics, 67th edition edn, CRC
- 1010 Press Inc, Boca Raton, FL, USA. 2424 p.
- 1011 [Zhao et al. (2006)] Zhao, B., Zhu, L. and Yang, K.: 2006, Solubilization of DNAPLs by mixed
- 1012 surfactant: Reduction in partitioning losses of nonionic surfactant, *Chemosphere* **62**(5), 772–779.
- [Zhong et al. (2003)] Zhong, L., Mayer, A. and Pope, G.: 2003, The effects of surfactant
 formulation on nonequilibrium NAPL solubilization, *Journal of Contaminant Hydrology* 60(1-2), 55–
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- 1019 Figure 1: Schematic of 1D cell
- Figure 2: Pressure-water saturation curves for 0.5 and 0.1 mm GB: a) drainage-imbibition withoutenhancement and b) imbibition without surfactant and with SDBS
- 1022 Figure 3: Comparison of a) Capillary number (N_{ca}), b) Bond number (N_B), and c) Total trapping
- 1023 number (N_T) for 0.5 and 0.1 mm GB
- 1024 Figure 4: Fitting the experimental permittivity values as function of water saturation with the CRIM
- 1025 model: a) 0.5 mm GB and b) 0.1 mm GB
- 1026 Figure 5: Change of water saturation and permittivity (measured and estimated) as a function of
- drainage-imbibition cycle for 0.5 and 0.1 mm GB in 1D cells (with and without enhancements)
- 1028 Figure 6: Fitting the experimental resistivity values as a function of water saturation with Archie's
- 1029 law: a) 0.5 mm GB and b) 0.1 mm GB $\,$
- 1030 Figure 7: Change of water saturation and resistivity (measured and estimated) as a function of
- drainage-imbibition cycle for 0.5 and 0.1 mm GB in 1D cells (with and without enhancements)
- 1032 Figure 8: Drainage-imbibition experiments in the Hele-Shaw cell (example with 0.5 mm GB)
- 1033 Figure 9: Fitting the experimental optical density values as a function of water saturation (without
- 1034 enhancement, with chemical and thermal enhancements)
- 1035 Table 1: Results of drainage-imbibition experiments for 0.5 and 0.1 mm GB

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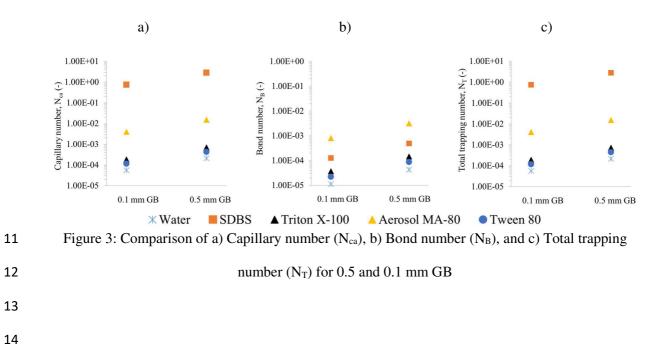


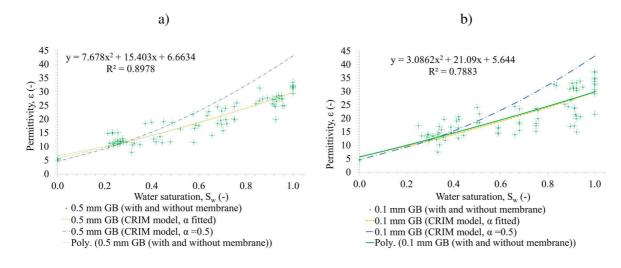
(D: drainage, I: imbibition, VG: Van Genuchten Mualem fitting values; Exp : raw values; Surf:

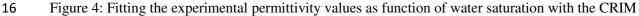
addition of SDBS)

Figure 2: Pressure-water saturation curves for 0.5 and 0.1 mm GB: a) drainage-imbibition without
enhancement and b) imbibition without surfactant and with SDBS

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model: a) 0.5 mm GB and b) 0.1 mm GB

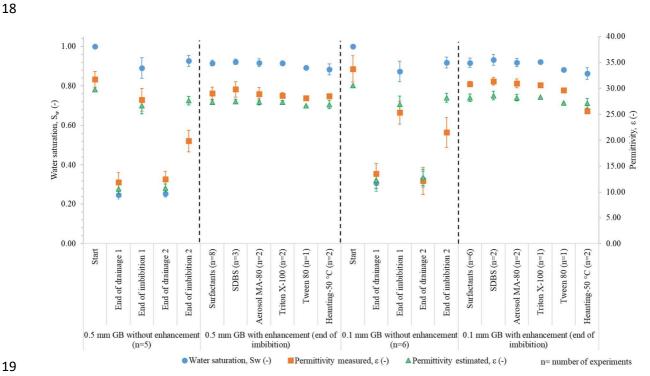
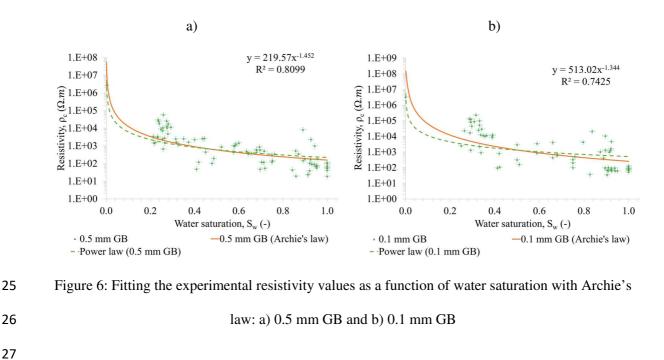




Figure 5: Change of water saturation and permittivity (measured and estimated) as a function of drainage-imbibition cycle for 0.5 and 0.1 mm GB in 1D cells (with and without enhancements)



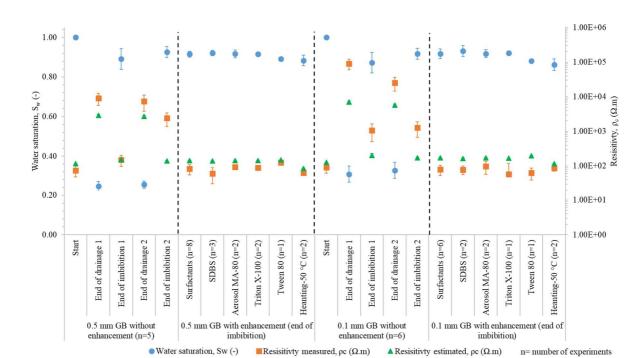
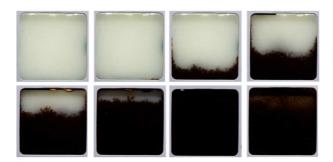




Figure 7: Change of water saturation and resistivity (measured and estimated) as a function of
drainage-imbibition cycle for 0.5 and 0.1 mm GB in 1D cells (with and without enhancements)

Drainage 1



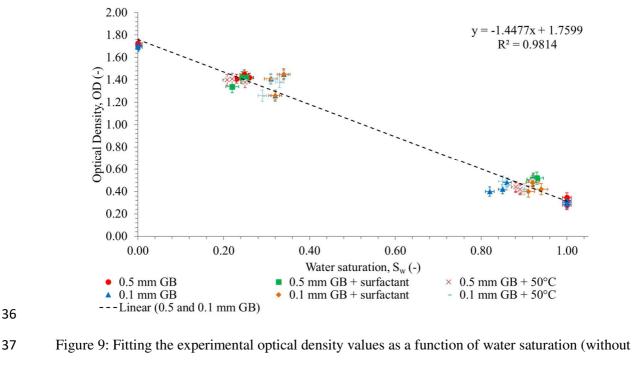
Imbibition 1



Drainage 2



33 Figure 8: Drainage-imbibition experiments in the Hele-Shaw cell (example with 0.5 mm GB)





enhancement, with chemical and thermal enhancements)

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Table 1: Results of drainage-imbibition experiments for 0.5 and 0.1 mm GB

Parameters	0.5 mm GB		0.1 mm GB	
T difunctions	Drainage D	Imbibition I	Drainage D	Imbibition I
α (m ⁻¹)	26.06	38.36	14.72	23.29
n (-)	15.35	5.15	9.49	16.98
SSE	0.0038	0.0025	0.0056	0.0214
S _{rn} (-)	0.000	0.109	0.000	0.127
S _{rw} (-)	0.248	0.248	0.309	0.309

Graphical abstract

